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Environmental Sciences Ltd.

Callander Bay Subwatershed Phosphorus Budget

Prepared By: Hutchinson Environmental Sciences Ltd.
Prepared For: North Bay-Mattawa Conservation Authority

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Ms. Sue Miller
North Bay – Mattawa Conservation Authority
15 Janey Avenue
North Bay, Ontario
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Dear Ms. Miller:

Re: Callander Bay Subwatershed Phosphorus Budget – Final Report

It is with pleasure that I submit this final report of the Callander Bay Subwatershed Phosphorus Budget. The report includes several revisions to the November 2nd, 2010 draft report to address comments received by the NBMCA, the TAC, and the public at the Public Meeting in January 2010. Revisions were also made with updated land cover and septic system data received from the NBMCA in December 2010.

I thank the NBMCA for giving us the opportunity to work on this assignment and for providing continued guidance and technical support. I also thank members of the TAC for their dedicated input to the project.

I would be happy to discuss any of the details of the report with you and to provide further support to the NBMCA in their future lake and watershed management endeavours.

Sincerely,
HESL




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Appendices

Appendix A. Data and Phosphorus Budget Calculations (digital)



1. Introduction

In recent years, there has been growing concern over the occurrence of algal blooms dominated by *Cyanobacteria* species (commonly known as bluegreen algae) in Callander Bay as well as Wasi Lake, a shallow productive lake in the Callander Bay watershed. Callander Bay is the municipal drinking water source for the Town of Callander. Both Callander Bay and Wasi Lake have elevated concentrations of the algal nutrient phosphorus ($>20 \mu\text{g/L}$), which is known to increase algal biomass and the risk of cyanobacterial blooms.

Cyanobacterial blooms not only affect the aesthetic and recreational water quality of water bodies (e.g., by decreasing water clarity or causing taste and odour problems), but many species of cyanobacteria produce toxins that can pose a risk to human health through ingestion or direct contact. The human health risk is particularly important for the municipal water supply because of cyanobacteria blooms dominated by taxa that are known to produce toxins. The cyanotoxin ‘microcystin’ is listed as a drinking water issue for the Callander intake under Ontario’s Drinking Water Source Protection (DWSP) program as mandated by the *Clean Water Act* (2006) (NBMCA, 2010).

Callander Bay is a large bay with a surface area of $\sim 1,206$ ha located at the east end of Lake Nipissing. It has a comparatively large catchment area ($\sim 28,413$ ha) that receives overland runoff and discharge from tributaries draining the Wistiwasing (Wasi) River subwatershed, and portions of the Bear-Boleau Creeks and La Vase River subwatersheds (Figure 1). Dominant flows in Callander Bay are towards the main body of Lake Nipissing although the bay likely mixes partially with water from the lake during periods of wind-induced flow reversals (HESL, 2010).

Callander Bay and Wasi Lake have been subjected to a variety of human disturbances that increased the supply of phosphorus since settlement of the area in the mid 1800s. These include logging and sawmill operations, agricultural activities and rural and urban development. Notably, operation of the Portage Dam at the outlet of Lake Nipissing following its construction in 1950 resulted in significant lowering of lake water levels, particularly in spring, which in turn is thought to have altered flushing rates and mixing regimes in Callander Bay. A study of historical changes of total phosphorus concentrations using paleolimnological techniques showed that these hydrological changes were coincident with increased phosphorus concentrations in Callander Bay which may have been a result of altered internal nutrient dynamics (AECOM, 2010). Since ~ 1950 , the combination of human sources and hydrological changes were inferred to have increased total phosphorus concentrations from moderate levels (~ 15 to $20 \mu\text{g/L}$) indicative of ‘mesotrophic’ conditions, to greater than $20 \mu\text{g/L}$ indicative of ‘eutrophic’ conditions.

Despite large-scale human influences over the past century there is no evidence, from either the paleolimnological study or monitoring data, of increased total phosphorus concentrations in Callander Bay over the past few decades (AECOM, 2009; HESL, 2010) that would cause the recent increase in cyanobacterial bloom activity. For many water bodies with elevated phosphorus concentrations like Callander Bay and Wasi Lake (including those with minimal watershed disturbance or human nutrient loading), however, effects of recent warming with climate change can exacerbate cyanobacterial blooms (e.g., AECOM, 2009; Wagner and



Adrian, 2009; Rühland et al., 2010). It is therefore conceivable that Callander Bay and Wasi Lake, given their pre-existing total phosphorus concentrations that increase the risk of nuisance algal growth, have enhanced susceptibility to climate-mediated cyanobacterial blooms.

Development of effective watershed management plans to reduce the risk of cyanobacterial blooms in Callander Bay and Wasi Lake requires control of human sources of phosphorus and thus knowledge of the relative contribution of all sources of phosphorus in the watershed – that is, a phosphorus budget. A phosphorus budget is also required to evaluate and further inform aspects of DWSP for the Callander drinking water intake, specifically:

- 1) delineation of areas within the defined Vulnerable Area of the intake that contribute phosphorus to the bay, (the ‘issue contributing area’),
- 2) identification and designation of activities in the issue contributing area that are prescribed to be drinking water threats and that may result in the release of phosphorus as ‘significant drinking water threats’, and
- 3) development of policies to mitigate these significant drinking water threats related to phosphorus.

The following study provides a subwatershed scale phosphorus budget for Callander Bay and Wasi Lake that quantifies human and natural sources of phosphorus using a combination of export coefficient modelling and measured phosphorus loadings. Recommendations for future monitoring requirements and potential mitigation strategies for the most relevant sources of human phosphorus are provided to guide future management plans. Finally, results of the phosphorus budget are assessed as they relate to DWSP for the Callander drinking water intake.

2. System Characterization

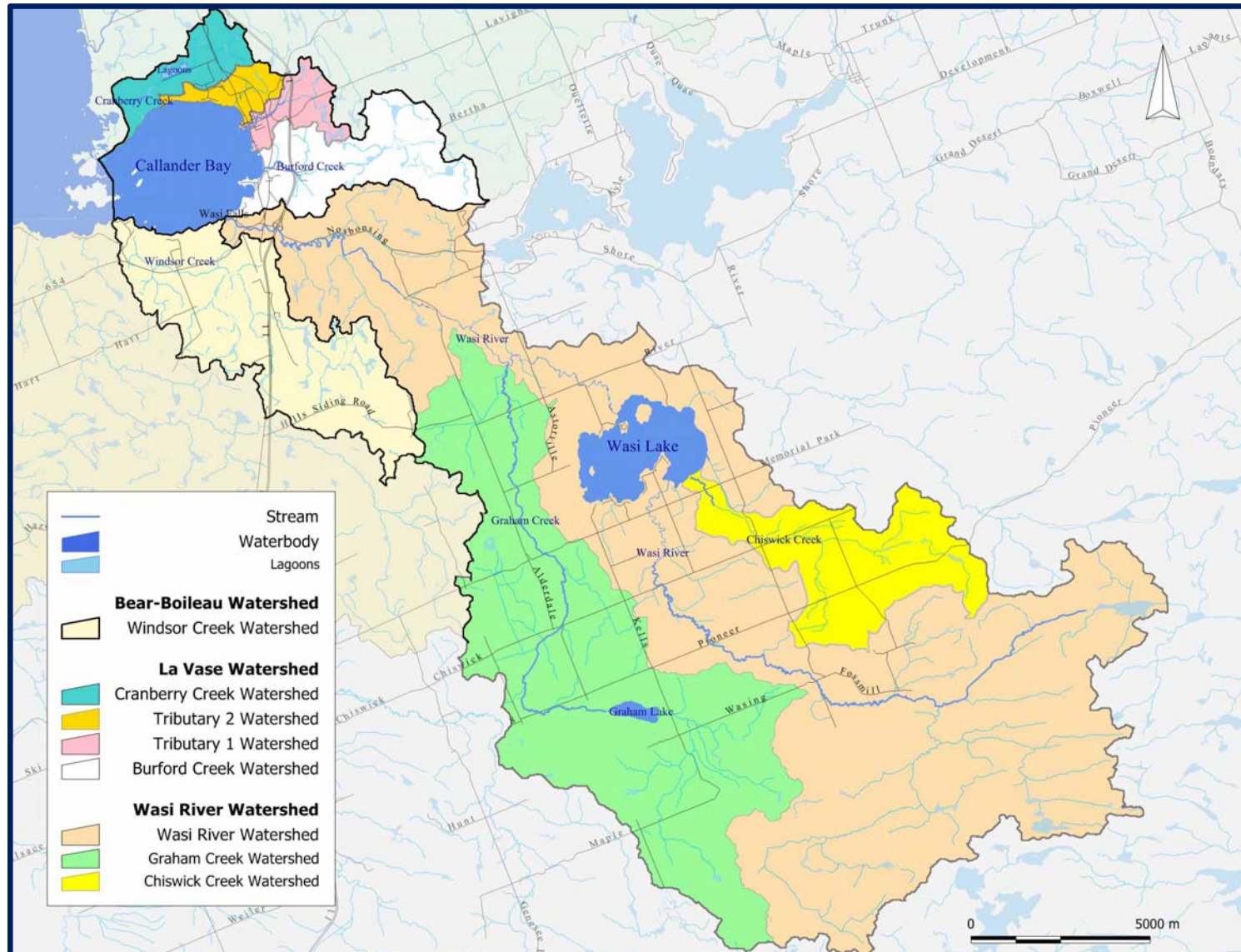
Phosphorus is plentiful in the natural environment as it occurs in inorganic phosphate-bearing rocks (e.g., apatite) and soils, and as organic phosphates in all living cells. The supply of phosphorus to surface water is dependent upon hydrological conditions, atmospheric sources, land cover, geology, soil type and depth, human activities, and climate. The following sections provide an overview of the primary factors influencing phosphorus loading to Callander Bay and Wasi Lake and resultant observed phosphorus concentrations in these water bodies as they relate to the development and validation of the phosphorus budget.

2.1 Hydrology

Phosphorus is most readily transported by water. It enters surface water from precipitation, tributary flow, overland runoff and groundwater. In surface water bodies, the concentration of phosphorus is then moderated by the quantity of water in the surface water body, mixing regimes that influence loss to the sediments and outflow which transports water carrying phosphorus downstream. Hydrological characteristics of a water body and its watershed are therefore important in the determination of phosphorus loads and concentrations in surface water.



Figure 1. Callander Bay and watershed area.



The Wasi River is the largest subwatershed of Callander Bay. It drains an area of ~23,433 ha, which represents ~82% of the Callander Bay catchment. Flows in the Wasi River have been monitored by the Water Survey of Canada (WSC) since 2008 (<http://www.wsc.ec.gc.ca/hydat/H2O>), providing two full years of data (Table 1). Tributary flows, however, can vary considerably from year to year depending on precipitation patterns; the total annual discharge of the Wasi River differed by more than 20% between 2008 and 2009 (Figure 2). Flows in these two years may therefore not be representative of the long-term average conditions. The available flow data from 2010 are still preliminary and should be viewed with caution until they have been verified by the WSC, but 2010 monthly flows to date are much lower than those observed in 2008 and 2009 (Figure 2).

Table 1. Water Survey of Canada Hydrometric Station Summary

Station Name	Wasi River near Astorville	La Vase River at North Bay
Station ID	02DD024	02DD013
Data Years	2007-2009	1974-2009
Measurement Type	Flow and Level	Flow and Level
Gross Drainage Area (km ²)	235 ¹	70.40
Geographic Coordinates	46°10'41" N 79°18'36" W	46°15'48" N 79°23'42" W

The nearby La Vase River flows have been monitored continuously since 1974 (Table 1) and can be used to assess the representativeness of the Wasi River flow data. While the La Vase River subwatershed is considerably smaller than the Wasi River subwatershed, flows in both rivers displayed similar seasonal patterns in 2008 and 2009 (Figure 2) and depth of runoff (flow per unit drainage area) in the two subwatersheds is similar (Table 2), suggesting similar hydrological characteristics.

Like most areas of north eastern Ontario, flows in the La Vase River were elevated throughout the summer months in 2008 (Figure 2) resulting in a total annual discharge of 6.4×10^6 m³ in comparison to the long-term 34-year mean of 4.6×10^6 m³. This reflects the higher than average precipitation in 2008. Total precipitation of 1335 mm in 2008 (Figure 3) was 23% greater than the 2007-2009 average of 1161 mm and 27% greater than the 1971-2000 average of 1007 mm, measured at the North Bay airport. The overall mean monthly flows in the La Vase River in 2009 were similar to the long-term means (Figure 2). It is therefore likely that the 2008 Wasi River flows also represent higher than average flow conditions and that the 2009 flows more closely resemble the long-term mean conditions for this river. This is supported by the fact that the runoff depth of 0.43m calculated from the Wasi River discharge data for 2009 is equivalent to the average annual runoff depth of 0.44m calculated for the La Vase River by Gartner Lee Ltd. (GLL, 2008) using a hydrological modelling approach (Table 2). The 2008 runoff depth of 0.53m represents a 23% increase over mean annual runoff.

Additional flow data were collected by the NBMCA from 4 tributaries draining to Callander Bay from May to August in 2009 (Table 3). Depth of runoff for the catchment areas of these tributaries for this 4-month period ranged from 0.07 m for Windsor Creek to 0.14 m for Cranberry Creek, which drains Cranberry Marsh and contains flows from discharge of the Callander waste water treatment plant lagoons in spring and fall. Runoff depth for the

¹ Confirmed as 235 km² by NBMCA, WSC database provides a value of 301 km², which is larger than the entire Callander Bay watershed and was therefore considered to be in error.



catchments of tributary NB-323 (0.10 m) and Burford Creek (0.09 m), both located within the greater La Vase River subwatershed, were similar to the runoff calculated for the entire La Vase River subwatershed (0.11 m) over the same period suggesting that localized differences in runoff within these areas of the watershed are minimal during the summer months.

Figure 2. Wasi River and La Vase River monthly stream discharge (2008-2010).

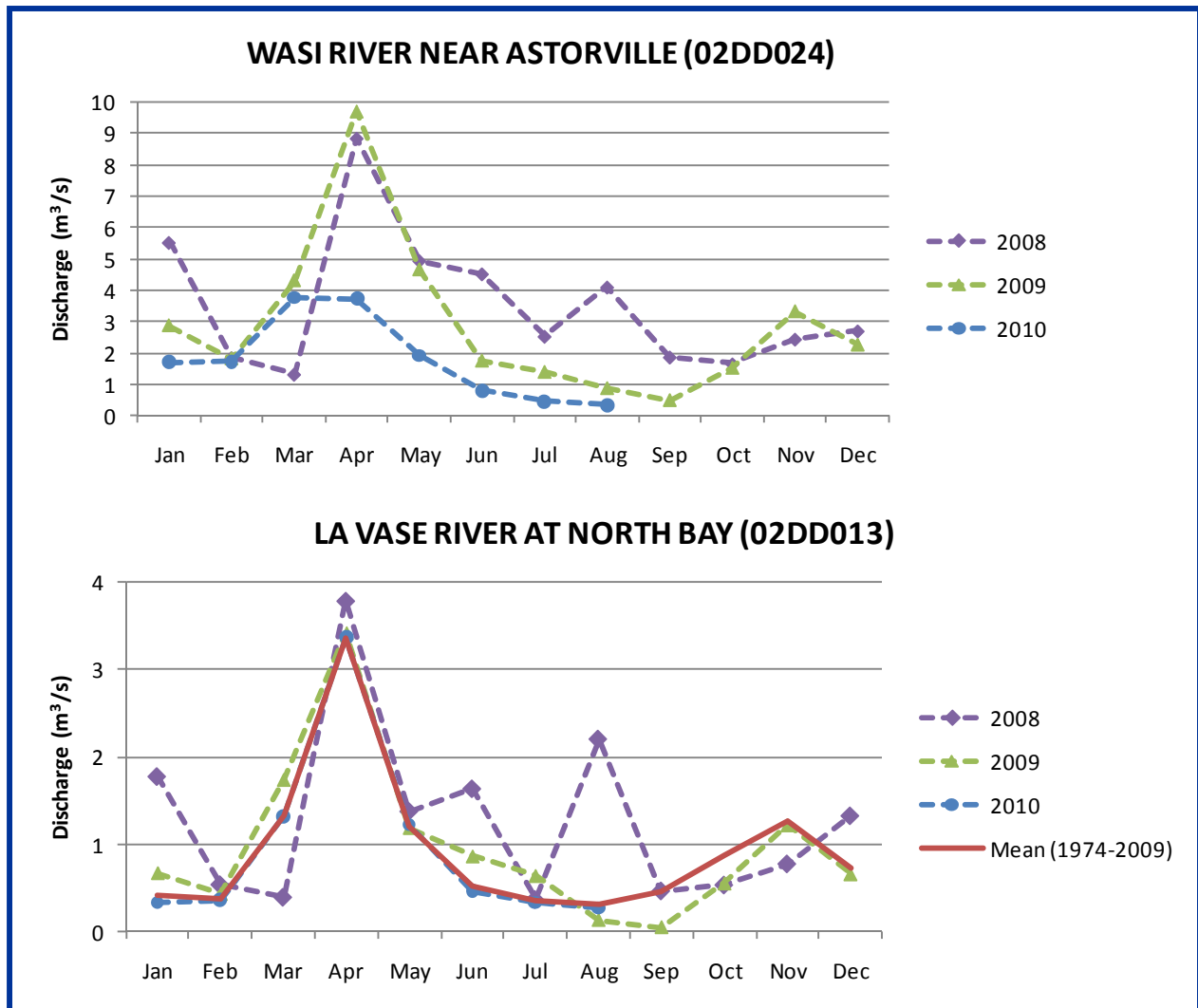


Figure 3. Precipitation at North Bay Airport.

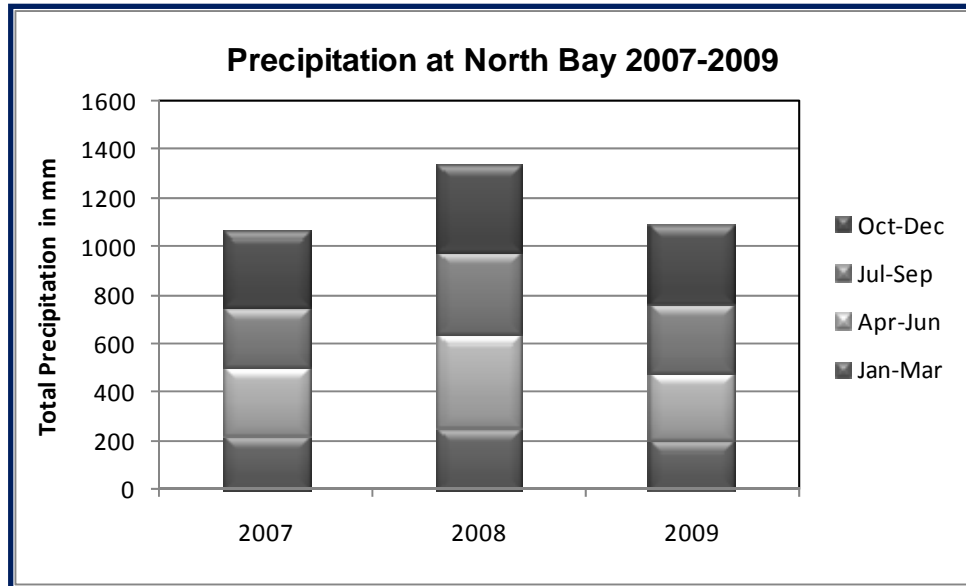


Table 2. Wasi River and La Vase River Discharge and Depth of Runoff Summary

Year	Discharge at Monitoring Gauge ($m^3 \times 10^6$)	Depth of Runoff in Subwatershed (m)	Area of Subwatershed Draining to Callander Bay (km^2)	Total Discharge to Callander Bay ($m^3 \times 10^6$)
WASI RIVER				
2008	111.8	0.53	234.6	123.9
2009	91.3	0.43		101.3
LA VASE RIVER				
2008	40.8	0.58	11.06	6.4
2009	31.0	0.44		4.9
Mean (1974-2009)	29.4	0.42		4.6
Water Budget Calculation (Table 5.1, GLL, 2007)		0.438		

Table 3. Discharge and Depth of Runoff Summaries for Tributaries of Callander Bay, May to August 2009

Tributary	Mean Discharge (m^3/s)	Discharge Volume ($m^3 \times 10^6$)	Drainage Area (km^2)	Depth of Runoff (m)
Burford Creek	0.11	1.16	12.7	0.09
Cranberry Creek	0.04	0.4	2.8	0.14
NB-323	0.03	0.34	3.4	0.10
Windsor Creek	0.16	1.7	25.8	0.07
La Vase River	0.71	7.5	70.4	0.11
Wasi River	2.15	22.8	235	0.11



Based on an estimated mean annual runoff of 0.42 m calculated from measured flows in the Wasi and La Vase Rivers, approximately $124 \times 10^6 \text{ m}^3$ of water is discharged from Callander Bay to Lake Nipissing each year. Inflow from Lake Nipissing can occur periodically in response to wind events that cause flow reversals back into the bay (Northland Engineering, 1993). Flow reversals are likely of short duration and have no effect on the net total annual discharge from Callander Bay to Lake Nipissing. The volume of Callander Bay is approximately $66 \times 10^6 \text{ m}^3$ (assuming a mean depth of 5.5 m) and so at this discharge rate, the total volume of water in Callander Bay is replaced approximately 1.9 times per year under average conditions by runoff from the watershed.

2.2 Land Cover and Land Use

The amount of phosphorus supplied to a water body by runoff is dependent on the type of land cover and land use. Phosphorus is exported from natural undeveloped areas (e.g., forest, wetland, grassland) from the decomposition of plant material, animal waste and from soil erosion. Human activities and disturbance of natural areas in the watershed typically increase the supply of phosphorus in runoff, for example by increasing erosion, fertilizer and detergent use, and human waste disposal.

Land cover in the Callander Bay watershed was classified into 12 classes using QuickBird satellite imagery and verification by ground-truthing (Figure 4, from Liscombe 2009). The imagery was captured between May and June 2007 with a spectral resolution of 0.6m but this was reduced to 2.8m for data processing. For the purposes of the phosphorus budget, the 12 QuickBird land classes were grouped into 7 classes (i.e., agriculture, forest, wetland, grassland, bare rock, urban (as infrastructure) and open water). Areas of each class were then determined on a subwatershed scale and for the Intake Protection Zones (IPZs) of the Callander drinking water intake (Figure 5).

A relatively large area (~10% of the subwatershed) was classified as cloud cover/shadow in the QuickBird classification and so specific land uses for these areas could not be determined. The area of shadow/cloud within each subwatershed was therefore added in proportion to the known land cover classes (e.g., if forest accounted for 30% of the land area, 30% of the cloud cover/shadow area was assigned to the forest class).

The spectral resolution of the QuickBird imagery does not allow differentiation of the types of agricultural lands (e.g., pasture and row crops) within the agriculture class or within manicured grassy areas (e.g., lawns and golf courses) that have markedly different phosphorus export. Therefore, one coefficient was assigned to all of these classifications which represented an average value for agricultural land uses (see Section 4.2.2).

Major land cover classes are summarized by area (Table 4) and by percent cover (Figure 6) for the Callander Bay watershed. Land cover is predominantly natural with forest (16,624 ha or 59%), wetland (6,335 ha or 22%) and grassland (417 ha or 1%) comprising 82% (23,376ha) of the total watershed area of 28,413 ha (excluding the surface area of Callander Bay). Agricultural areas (including large parkland areas and golf courses) total 3,280 ha or 12%, of the watershed, bare rock makes up less than 1 ha and urban areas (buildings, paved areas, urban lawns) 388 ha (1%). Open water makes up the remaining 9% of the watershed.



The total area of human disturbance is 3,668 ha, most of which (3,280 ha or 89%) is agriculture. Larger parkland areas and golf courses are concentrated in the vicinity of the Town of Callander.

The proximity of disturbed areas to the Intake Protection Zones of the Callander drinking water intake has implications for the supply of phosphorus to Callander Bay. Phosphorus in overland runoff within the IPZs is more likely to reach surface waters and be transported to Callander Bay where it could promote cyanobacterial blooms. Therefore, identification of disturbed lands within the IPZs will allow any required remedial actions to be focussed where they are most likely to improve water quality in Callander Bay.

The Intake Protection Zones of the Callander drinking water intake encompass 45% of the land area in the Callander Bay watershed and include 46% of the disturbed lands, that is 44% of the agricultural area and 58% of the urban area in the Callander Bay watershed is located within 120-m of a water course or water body.



Figure 4. QuickBird-derived land classes in the Callander Bay watershed.

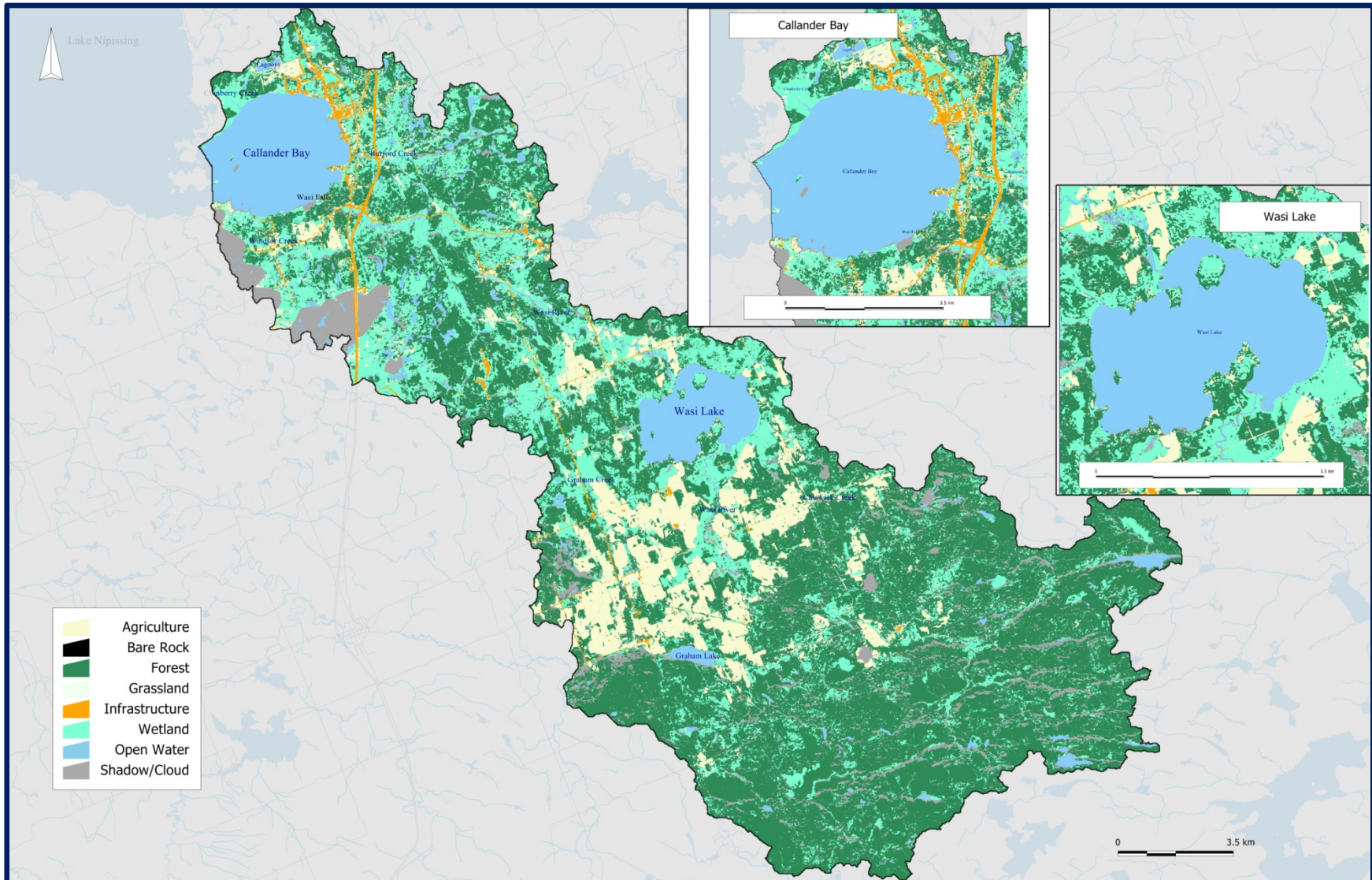


Figure 5. Intake Protection Zones (IPZs) of the Callander drinking water intake (from HESL 2010).

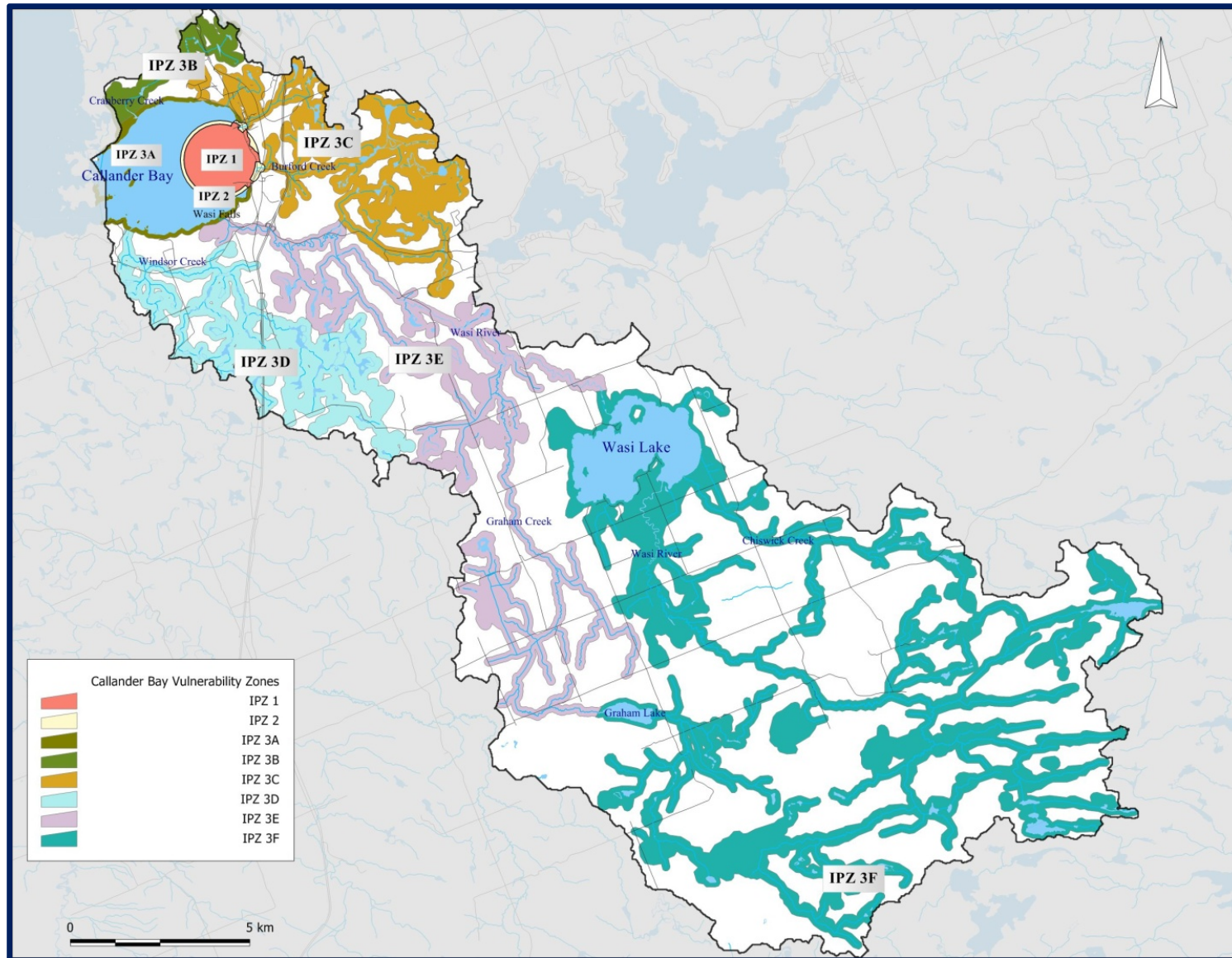


Table 4. Land Cover Summary by Callander Bay Subwatershed and Intake Protection Zone

Area		Land Cover (ha)						Total
		Forest ¹	Wetland ²	Grassland	Agriculture ³	Urban	Open Water	
By Subwatershed								
Wasi River Subwatershed	Wasi River catchment upstream of Wasi Lake	7,222	1,214	91	695	11	168	9,399
	Chiswick Creek catchment	1,365	275	35	187	3	53	1,918
	Total Wasi Lake and catchment	9,218	2,050	139	1,160	24	950	13,542
	Graham Creek catchment	3,800	1,242	68	1,436	58	111	6,716
	Total Wasi River subwatershed	14,582	4,462	279	2,829	161	1,121	23,433
La Vase River Subwatershed	Burford Creek catchment	615	470	33	71	41	64	1,295
	Tributary 1 catchment	128	109	8	42	39	17	343
	Tributary 2 catchment	84	78	12	30	40	8	252
	Cranberry Creek catchment	201	184	9	63	17	37	511
	Total La Vase River subwatershed	1,029	840	62	205	138	126	2,400
Bear-Boleau Creeks Subwatershed	Windsor Creek catchment	1,012	1,033	77	247	89	122	2,579
Callander Bay surface area							1,206	
Total Callander Bay watershed (includes surface area of bay)		16,624	6,335	417	3,280	388	2,574	29,619
By Intake Protection Zone								
IPZ-1		5	4	1	7	10	nd	27
IPZ-2		4	4	1	4	7	nd	20
IPZ-3a		41	63	6	24	28	nd	162
IPZ-3b		123	135	6	21	7	nd	292
IPZ-3c		724	582	41	117	76	nd	1,540
IPZ-3d		585	650	44	139	44	nd	1,461
IPZ-3e		922	855	38	549	42	nd	2,407
IPZ-3f		4,077	1,487	54	593	11	nd	6,222
Total Intake Protection Zones		6,481	3,779	192	1,455	224		12,131
% of Callander Bay Watershed		39	60	46	44	58		45⁴

Notes:

nd – no data

¹includes coniferous, deciduous, mixed forest and forest regeneration QuickBird classes

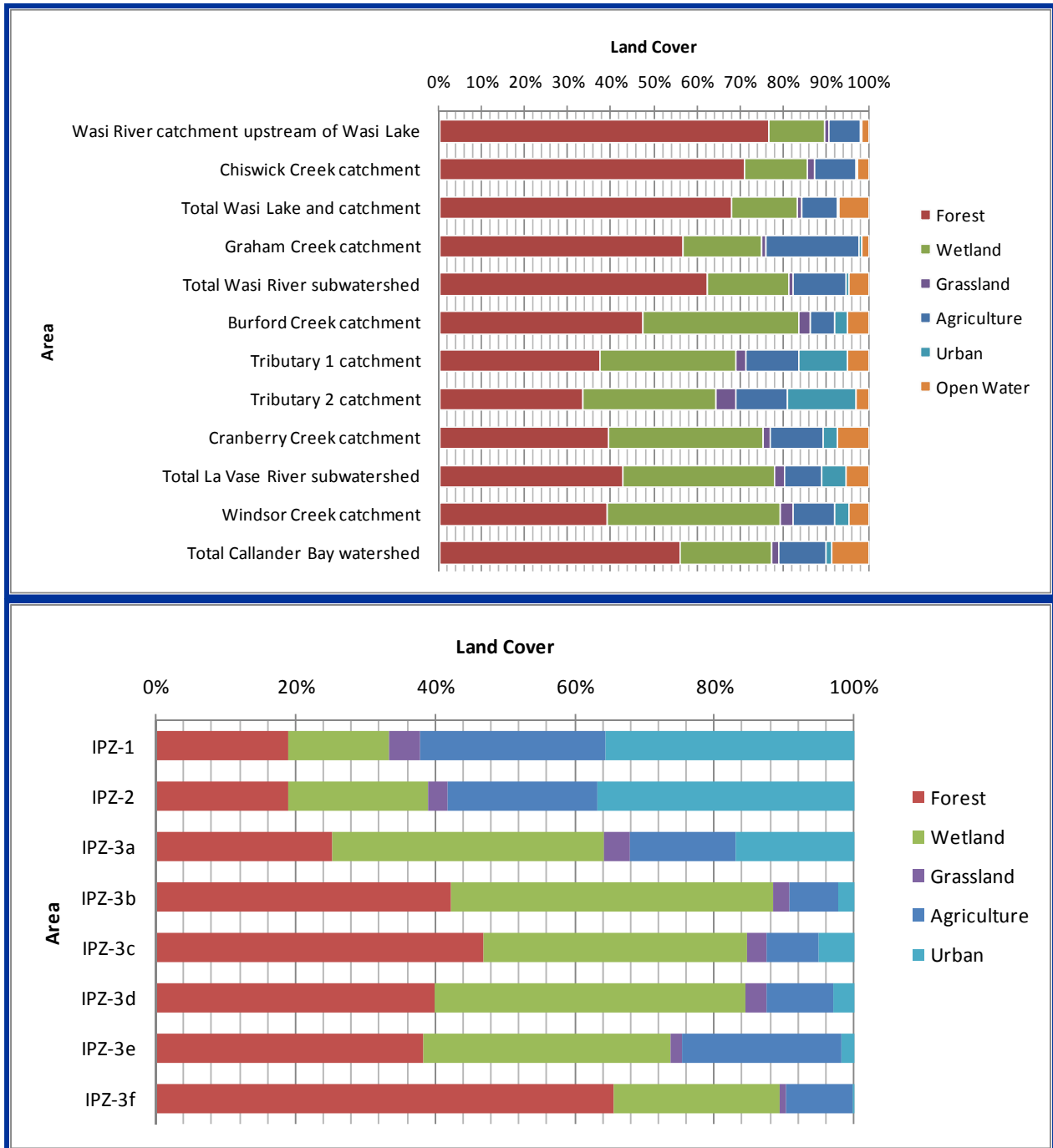
²includes marsh, swamp and treed fen QuickBird classes

³includes agriculture 1 and agriculture 2 QuickBird classes (which includes golf courses and some manicured lawn areas)

⁴calculated as percent of Callander Bay land area only



Figure 6. Percent land cover in the Callander Bay watershed by subwatershed area and Intake Protection Zone.



2.3 Phosphorus Concentrations

The concentration of phosphorus in surface water is a function of the supply of phosphorus and the input of water and is moderated by aspects of basin morphometry (e.g., depth, surface area) that influence internal phosphorus dynamics (e.g., mixing regimes, sedimentation and resuspension of phosphorus). Monitoring initiatives of the NBMCA and the Ministry of the Environment's Provincial Water Quality Monitoring Network (PWQMN) provide recent phosphorus concentration data for Callander Bay, Wasi Lake and Wasi River that is representative of existing conditions and can be used with confidence to develop and validate the phosphorus budget as described in Section 3.

The NBMCA has been monitoring total phosphorus concentrations as euphotic zone composite samples in Callander Bay and Wasi Lake on a nearly biweekly basis over the open water season since 2007. This monitoring program was expanded to include several locations along the Wasi River (11 sites), Graham Creek (6 sites) and Chiswick Creek (3 sites) from June to October of 2009 (Figure 7). In 2010, monitoring included an additional 2 sites, one on Burford Creek and one on Windsor Creek.

For the Wasi River, longer term total phosphorus concentration data are available from the Province's Provincial Water Quality Monitoring Network (PWQMN) station, which is the same location as the NBMCA's station W11. The Wasi River PWQMN station has been monitored sporadically since 1984, with a total of 14 years of total phosphorus concentration data collected monthly over the ice free period, typically from April to November.

Raw total phosphorus concentration data from all sources are presented in Appendix A (digital format) along with details of data evaluation (i.e., identification of outliers).

The mean ice free total phosphorus concentration in Callander Bay from 2007 to 2009 was 21.5 µg/L, which exceeds the Provincial Water Quality Objective (PWQO) of 20 µg/L for the protection against nuisance growth of algae in lakes. Concentrations are typically lowest in the spring (mean spring TP = 16.2 µg/L) and increase over the growing season reaching highest concentrations in mid to late summer (mean August TP = 24.3 µg/L) (Figure 8).



Figure 7. NBMCA phosphorus monitoring sites in Wasi River, Chiswick Creek and Graham Creek.

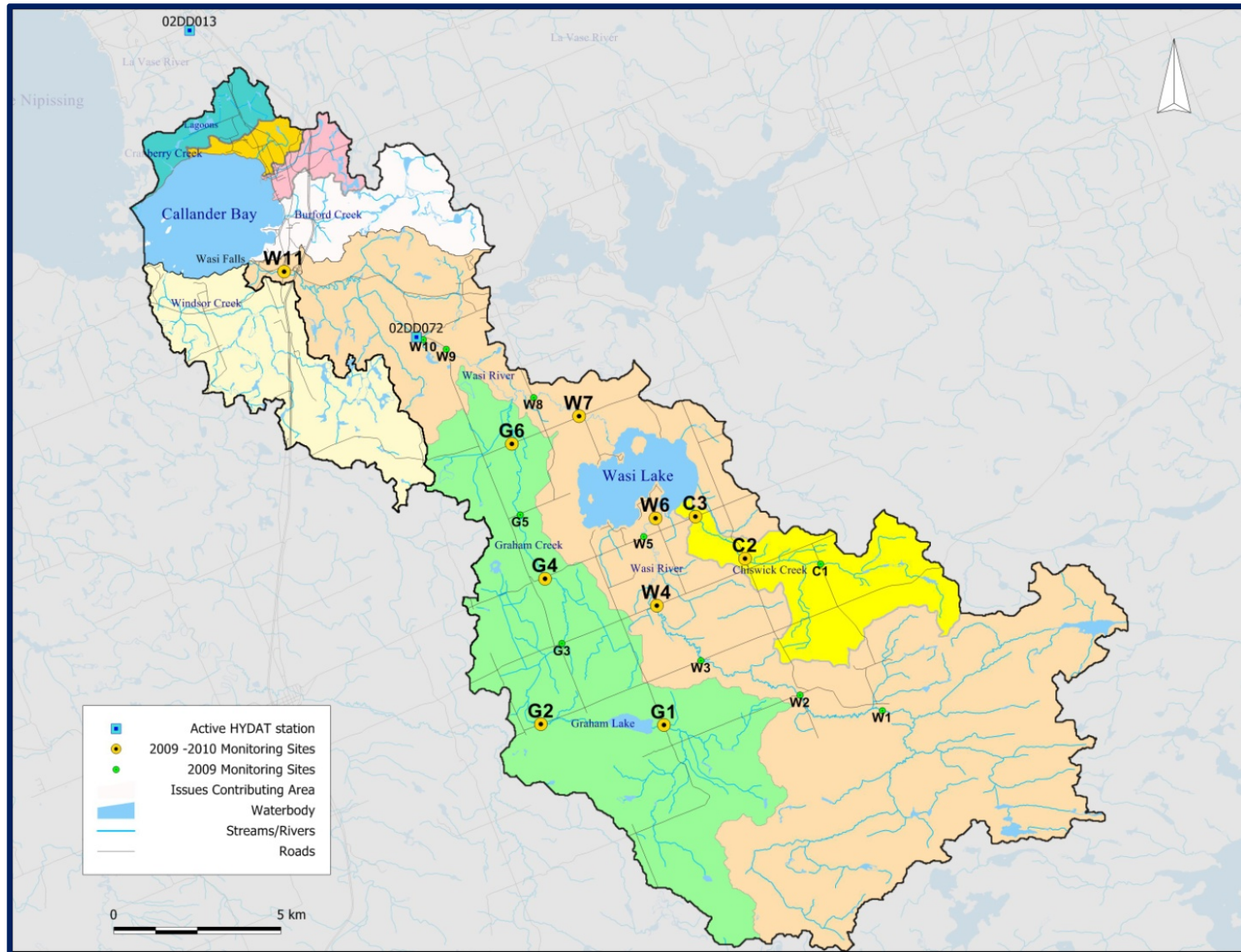
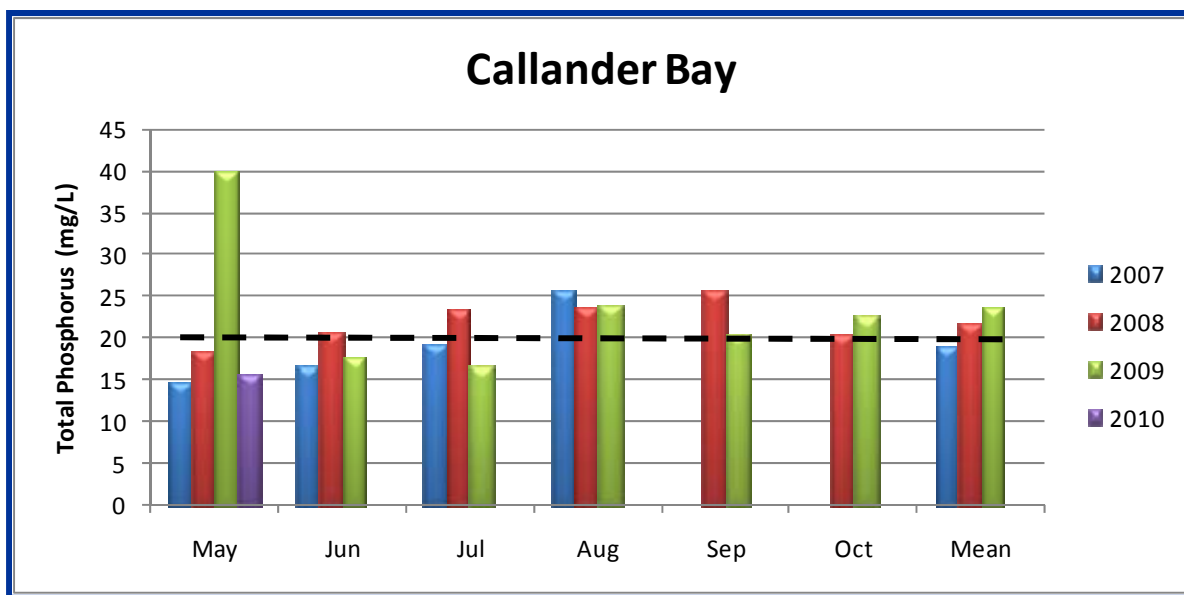


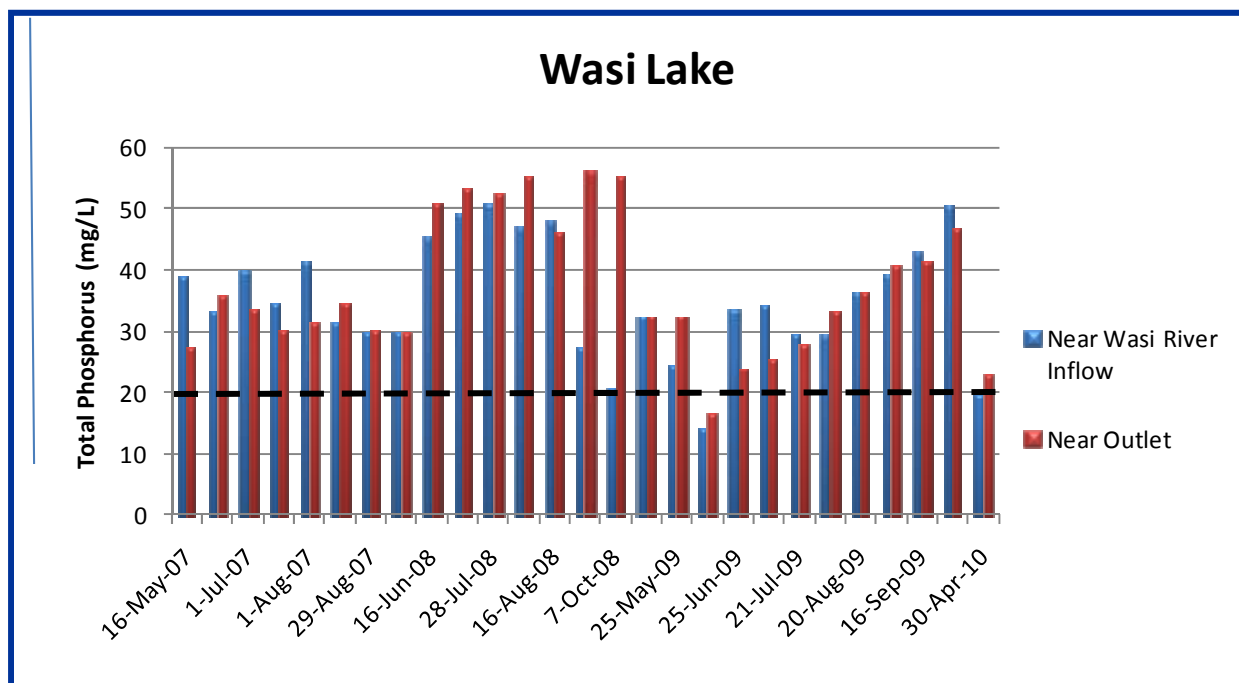
Figure 8. Mean monthly total phosphorus concentrations in Callander Bay (2007-2010) monitored by the NBMCA.



Note: Observed total phosphorus concentrations in May of 2009 are suspect of contamination or field/laboratory error. Dashed line indicates PWQO of 20 µg/L.

For Wasi Lake, ice free total phosphorus concentrations from 2007 to 2009 were very high. The mean concentration of 37.1 µg/L greatly increases the risk of cyanobacterial blooms. Unlike Callander Bay, patterns in phosphorus concentration between years and over the ice free period are highly variable (Figure 9). Mean ice free concentration ranged from 44.8 µg/L in 2008 to 32.9 µg/L in 2009, representing a 31% difference in concentration between years. In 2007 and 2008, phosphorus concentrations did not display any apparent trend over the growing season, but in 2009, a strong increasing trend occurred with concentrations increasing from 30.2 mg/L in May to 48.7 µg/L in October. Variability is also noted between phosphorus concentrations measured near the inlet of the Wasi River and the outlet of Wasi Lake. Variability in phosphorus concentrations in Wasi Lake indicates that this lake is highly sensitive to hydrological differences (e.g., river discharge and precipitation patterns) and variability in phosphorus loading or internal lake dynamics over the course of the growing season. Heightened sensitivity to changes in phosphorus and water loads may be due to the large surface area of the lake relative to the lake depth (mean depth ~2 m) in combination with the large watershed area of this lake relative to its surface area.

Figure 9. Mean monthly total phosphorus concentrations in Wasi Lake (2007-2010) monitored by the NBMCA.



Note: Dashed line indicates PWQO of 20 ug/L.

Total phosphorus concentrations in the Wasi River and tributaries in the Callander Bay watershed in 2009 and 2010 are summarized in Table 5 and Figures 10 to 12. Longer term monitoring data for the Wasi River from the PWQMN station are presented in Figures 13 and 14.

The Wasi River and smaller tributaries (Burford, Chiswick and Graham and Windsor creeks) monitored by the NBMCA all have mean total phosphorus concentrations above the PWQO of 30 µg/L for the protection against nuisance plant growth in rivers. For Chiswick and Graham creeks, phosphorus concentrations increase with distance downstream suggesting increased phosphorus loading to these tributaries as they become subject to human sources of phosphorus in the watershed (Figures 10 and 11). This trend is not consistently apparent along the length of the Wasi River (Figure 12); although concentrations typically increase downstream of station W2 with the occurrence of human disturbance in the watershed.

Mean annual phosphorus concentrations measured since 1985 in the Wasi River were variable, ranging from 27 µg/L to 42 µg/L with a mean concentration of 39 µg/L (Figure 13). This interannual variability likely reflects annual differences in precipitation patterns, but may also be due to differences in the timing of sample collection as concentrations can vary considerably over the ice free season (Figure 14). With the exception of 2004, mean monthly phosphorus concentrations in the Wasi River increased following snowmelt and the spring freshet reaching maximum concentrations in August, then decreased to near spring concentrations by fall. Regardless of interannual variability however, phosphorus concentrations in the river increased

significantly over the past 15 years (t-test, $p < 0.005$) from 36.6 $\mu\text{g/L}$ (1985-1989) to 42.1 $\mu\text{g/L}$ (2003-2010) (Figure 13).

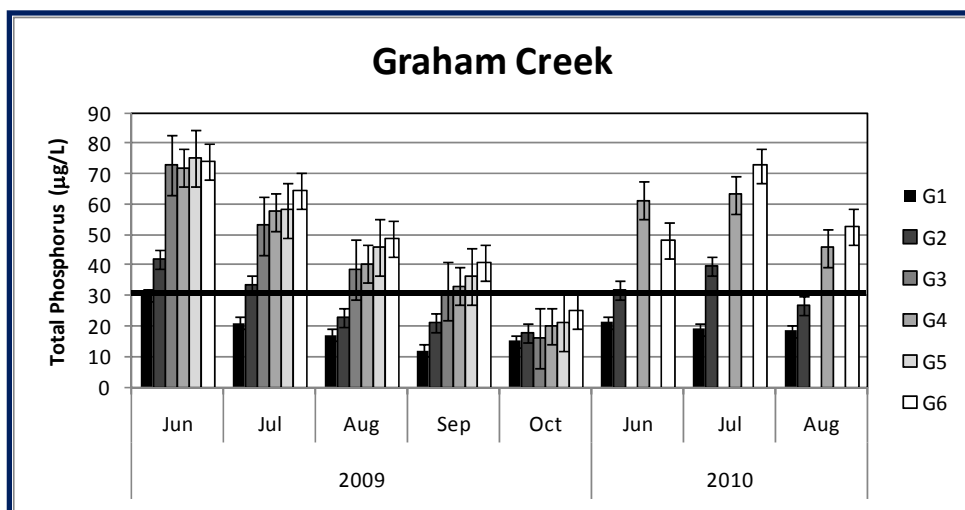
Table 5. Total Phosphorus Concentrations in Tributaries to Callander Bay, 2009-2010

Tributary (Code)	Total Phosphorus Concentration ($\mu\text{g/L}$)					n	
	Station	2009 ¹	2010 ²	Mean	Minimum		Maximum
BURFORD CREEK (BC)			37	37	22	47	5
BC1	nd	37	37	22	47	5	
CHISWICK CREEK (C)		46	34	40	10	93	39
C1	47	22	35	10	71	15	
C2	42	nd	42	21	93	9	
C3	49	46	48	23	73	15	
GRAHAM CREEK (G)		39	43	41	10	81	77
G1	18	20	19	10	30	14	
G2	28	34	31	18	52	15	
G3	43	nd	43	16	73	9	
G4	46	59	53	20	79	15	
G5	48	nd	48	21	75	9	
G6	52	57	55	25	81	15	
WASI RIVER (W)		46	48	47	14	100	130
W1	38	nd	38	28	49	8	
W2	44	29	37	14	72	13	
W3	42	nd	42	23	61	8	
W4	47	55	51	18	100	13	
W5	51	nd	51	36	98	8	
W6	50	56	53	35	84	13	
W7	44	44	44	31	57	13	
W8	44	nd	44	32	76	8	
W9	49	nd	49	28	87	8	
W10	51	53	52	32	81	14	
W11	46	47	47	28	62	24	
WINDSOR CREEK (WC)		nd	64	64	49	108	5
WC1	nd	64	64	49	108	5	

Notes: ¹ average of biweekly samples collected June to October, ² average of biweekly samples collected June to August, nd – no data

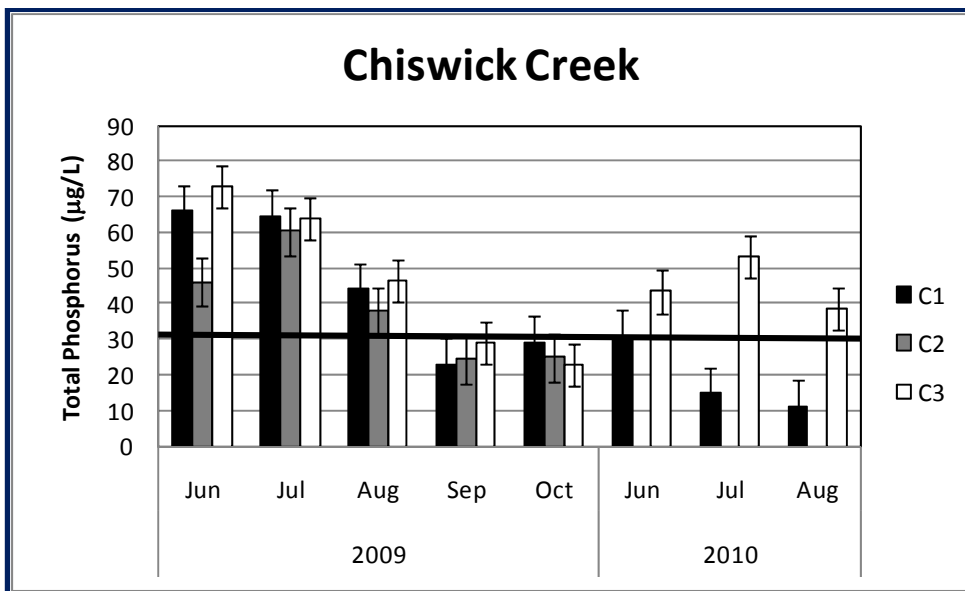


Figure 10. Mean monthly total phosphorus concentrations in Graham Creek (2009-2010). Sites numbers correspond to those provided in Figure 6 and increase with distance downstream in the river (G1 is the furthest upstream site and G6 is the furthest downstream site).



Note: Solid line indicates PWQO of 30 ug/L.

Figure 11. Mean monthly total phosphorus concentrations in Chiswick Creek (2009-2010). Sites numbers correspond to those provided in Figure 6 and increase with distance downstream in the river (C1 is the furthest upstream site and C3 is the furthest downstream site).



Note: Solid line indicates PWQO of 30 ug/L.



Figure 12. Mean monthly total phosphorus concentrations in Wasi River (2009-2010). Sites numbers correspond to those provided in Figure 6 and increase with distance downstream in the river. Sites 1-6 are located upstream of Wasi Lake and sites 7-11 are located downstream of Wasi Lake.

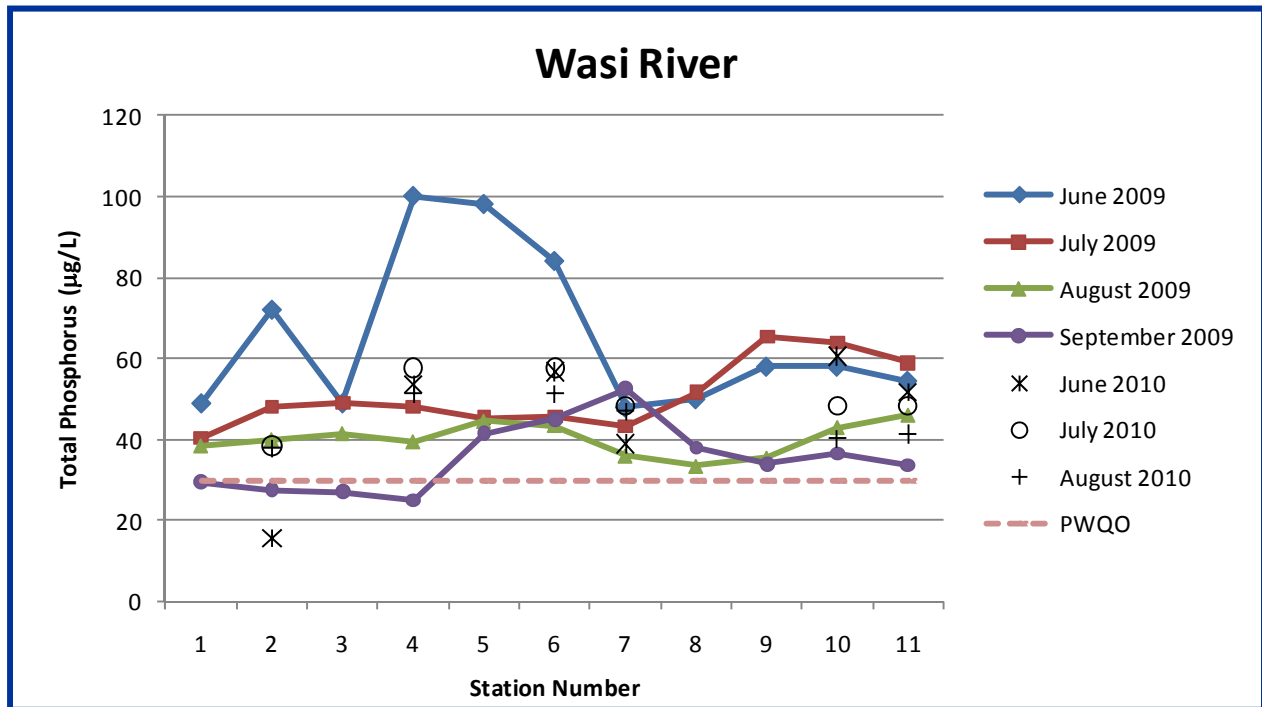
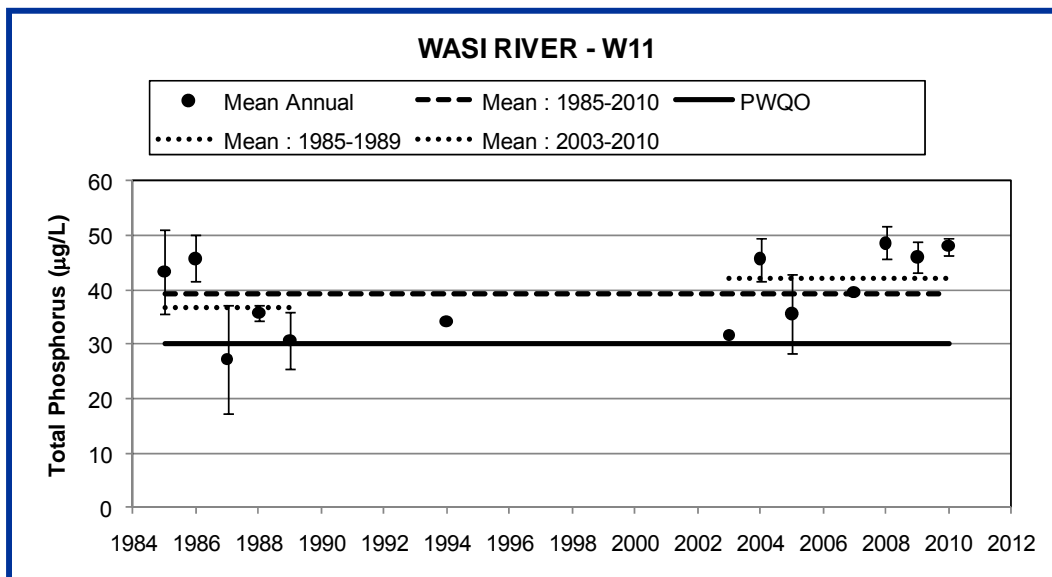
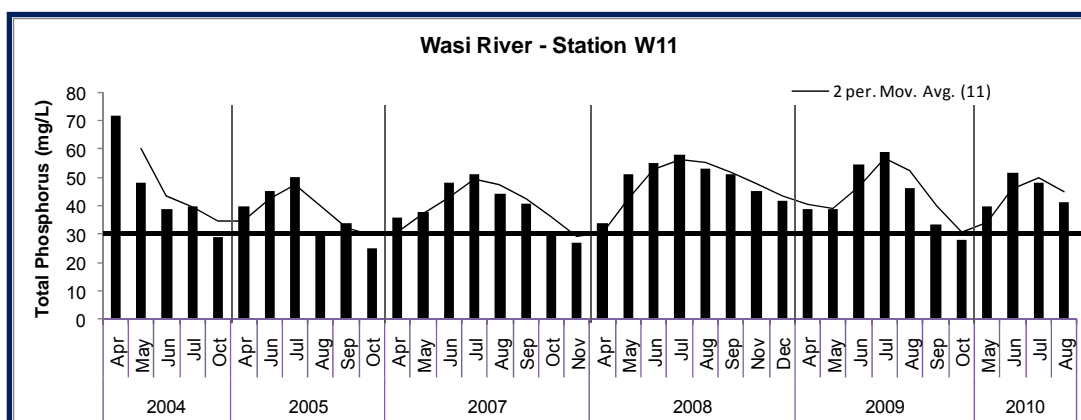


Figure 13. Mean annual ice-free total phosphorus concentrations at the mouth of the Wasi River (PWQMN station W11 data, 1985-2010).



Note: Includes years with at least 5 measured values in the ice-free season (April to October).

Figure 14. Mean monthly total phosphorus concentrations over the ice-free period in Wasi River at the PWQMN station (2004-2010).



Note: Solid horizontal line indicates PWQO of 30 ug/L.

3. Phosphorus Budget Approach

We used two approaches to estimate the loading of phosphorus to Callander Bay. Export coefficient modelling was used to estimate the loading from specific source areas in the watershed, to identify the most important source terms and, for the future, to allow assessment of how source-specific terms might change with specific management practices. Total annual loadings were calculated from measured data on flow and concentration in the Wasi River to compare with estimates made by export coefficient modelling. To further validate the loading calculations, observed phosphorus concentrations in Wasi Lake and Callander Bay were compared to concentrations predicted from a steady-state, mass balance water quality model.

3.1 Export Coefficient Modelling

An export coefficient modelling approach was used to calculate phosphorus loading to Callander Bay from non-point sources. The approach was developed in North America to predict nutrient inputs to lakes and streams (Dillon and Kirchner, 1975; Beaulac and Reckhow, 1982; Rast and Lee, 1983) and is now a well-established method of estimating phosphorus export when measured tributary flows and total phosphorus concentration data are lacking (e.g., Dillon *et al.* 1986, Johnes 1996, Winter and Duthie 2000, Paterson *et al.*, 2006). The export coefficient approach is also used where it is desirable to forecast nutrient export from a land area prior to a change in land use or prior to implementing Best Management Practices, in which case it is used as a predictive tool.

The use of phosphorus export coefficients for estimating phosphorus loading is based on the knowledge that specific land use types yield or export quantities of phosphorus to a downstream waterbody over an annual cycle. The export coefficients are developed from intensive, long-term monitoring programs that are typically carried out by academic institutions or government agencies. Knowing the area of land in a watershed devoted to specific uses and the quantities of nutrients exported per unit area of these uses (nutrient export coefficients), it is possible to

estimate total annual loads of phosphorus to Callander Bay from non-point sources using the equation:

$$L = \sum EiAi + P + S,$$

where L is the phosphorus load delivered to Callander Bay, Ei is the export coefficient selected for the specific land use/cover, Ai is the area of the land use/cover, P is the input from precipitation to the bay and S is the estimated input from septic systems within 300 m of the shoreline of Callander Bay and its tributaries. Export coefficients are expressed as rates (kg/ha/yr) and were derived from a survey of the literature.

Details of export coefficient selection and load calculations specific to each land use/cover type, precipitation and septic systems are provided in Section 4.

3.2 Measured Phosphorus Load Calculations

Phosphorus loads can be calculated for any water source carrying phosphorus to a water body (e.g., tributary, sewage effluent) if the total volume of water discharged to the water body and its phosphorus concentration are known, where:

$$P \text{ Load} = \text{Discharge} \times TP \text{ Concentration}$$

For the Callander Bay phosphorus budget, phosphorus loads were calculated using measured data for the Wasi River and for the Callander Waste Water Treatment Plant.

3.2.1 Wasi River

Phosphorus loading from Wasi River to Callander Bay was calculated using measured discharge data in 2009 from the WSC monitoring site near Astorville and total phosphorus concentration data collected by the NBMCA and from the PWQMN Wasi River station (2004 to 2010).

Ideally, phosphorus loads should be calculated using long-term data to capture variability in discharge and phosphorus concentrations from year to year. For the Wasi River, however, only two full years of discharge data (2008 and 2009) are available from the WSC monitoring site near Astorville. The 2009 discharge data were likely representative of the long-term average while the 2008 flows were higher than average, based on comparison with long-term flows in the LaVase River (Section 2.1, Table 2). Calculations are therefore based on 2009 discharge data only to better reflect average flow conditions.

Monthly runoff was calculated by dividing the monthly discharge of the river by the watershed area upstream of the monitoring station. Discharge from the total Wasi River watershed was then calculated for each month on a *pro rata* basis and summed to provide the total annual discharge.

Mean monthly total phosphorus concentrations were determined for the months of April to October using all available data from 2004 to 2010. Concentrations for the months of



November to May were estimated by taking the average concentration observed for the months of October and April.

3.2.2 Callander Waste Water Treatment Plant (WWTP)

Treated sewage discharge flows and phosphorus concentrations from the Callander Waste Water Treatment Plant (WWTP) were obtained from Annual Reports prepared by the Ontario Clean Water Agency (OCWA) for the Municipality of Callander (2002-2009). Phosphorus loads during the lagoon release times (spring and fall) were calculated by multiplying the total volume of effluent discharged by the mean total phosphorus concentration of the treated effluent. The annual total phosphorus load from the WWTP was the sum of the spring and fall loads.

3.3 Phosphorus Concentration Modelling

A variant of MOE's Lakeshore Capacity Model (LCM) was used to predict phosphorus concentrations in Wasi Lake and Callander Bay. This model is a steady-state, phosphorus mass balance model initially developed by Dillon et al. (1986) and revised in recent years by Hutchinson (2002) and Paterson et al. (2006), and is the recommended tool for setting development guidelines for Precambrian Shield lakes in Ontario (MOE, 2010). The model predicts total phosphorus concentration [TP] in lakes by estimating hydrologic and phosphorus loads from natural (watershed runoff and atmospheric deposition) and human (septic systems and land disturbance) sources and linking them together with an understanding of lake dynamics, using the equation:

$$[TP]_{ice\ free} = L_T * (1-R) * (0.965 * q_s)^{-1}$$

where L_T is the areal phosphorus loading rate (i.e., the total phosphorus loading divided by the lake surface area), R is the retention coefficient² and q_s is the areal water load (i.e., outflow discharge divided by lake surface area).

For Callander Bay and Wasi Lake, total phosphorus loading was estimated as the sum of all loadings from natural and human sources as calculated using the export coefficient approach described above. Hydrological data from the Wasi River were used to estimate water loads to Callander Bay from the entire watershed by areal pro-rating. The model used was a 'closed basin' model and therefore did not account for mixing of water from Lake Nipissing into Callander Bay. This mixing is dependent on wind events, would increase the flushing of water and therefore reduce the average phosphorus concentrations measured in the Bay. As a result, the water quality model would over-estimate phosphorus concentrations in Callander Bay, but not in Wasi Lake which does not mix with other water bodies.

Predicted phosphorus concentrations from the model were compared to measured concentrations to validate the phosphorus load calculations for Wasi Lake and to assess the response of Callander Bay. Predicted and measured concentrations that differed by 20% or less, provided confidence in the choice of export coefficients and the calculation of the phosphorus budget.

² The annual retention (R) is calculated as a function of the areal water load and a phosphorus settling velocity for either oxic or anoxic hypolimnion. Dillon et al. (1986)



Finally, the phosphorus concentration model was used to assess the responses of Callander Bay and Wasi Lake to reduced phosphorus loads that could be achieved by mitigation.

4. Phosphorus Budget Components

4.1 Natural/Undeveloped Phosphorus Loading

4.1.1 Atmospheric Loading

In nature, phosphorus has almost no gaseous forms and so the major transport mechanism is by water flow. Nevertheless, significant amounts of phosphorus are transported via the atmosphere as dust and dissolved in precipitation and deposited directly to the lake surface. For many lakes, atmospheric deposition constitutes a significant portion of the total phosphorus load, particularly for headwater lakes or other lakes with a large surface area relative to their catchment area.

Atmospheric loads are difficult to measure due to complexities with the collection and interpretation of precipitation chemistry data. It is preferable, therefore, to use estimates derived from regional, long-term study locations where reliable estimates of phosphorus in atmospheric deposition have been derived for multiple station datasets. In this case, a phosphorus deposition rate of 0.167 kg/ha/yr derived from 17-year records (1984-2001) at three meteorological stations in central Ontario at the Ontario Ministry of the Environment's Dorset Environmental Science Centre represents the nearest relevant atmospheric phosphorus deposition estimate (Paterson et al. 2006). This loading rate was multiplied by the surface areas of water bodies to calculate phosphorus loading to the surface of Callander Bay, Wasi Lake and open water areas in the watershed (Table 6). Atmospheric loads to land surfaces are captured by the export coefficients used to calculate watershed loads from land areas.

Atmospheric deposition contributes a phosphorus loading of 201 kg/yr directly to the surface of Callander Bay and 229 kg/yr to the surface of open water bodies in its watershed for a total loading of 430 kg/yr.



Table 6. Phosphorus Loading from Atmospheric Deposition to Open Water Areas

		Open Water Area	Phosphorus Loading
		(ha)	(kg/yr)
Area			
Wasi River Subwatershed	Wasi River catchment upstream of Wasi Lake	168	28
	Chiswick Creek catchment	53	9
	Total Wasi Lake and catchment (including Chiswick Creek and Wasi River catchments)	950	159
	Graham Creek catchment	111	19
	Total Wasi River subwatershed	1,121	187
La Vase River Subwatershed	Burford Creek catchment	64	11
	Tributary 1 catchment	17	3
	Tributary 2 catchment	8	1
	Cranberry Creek catchment	37	6
	Total La Vase River subwatershed	126	21
Bear-Boleau Creeks Subwatershed	Windsor Creek catchment	122	20
Callander Bay		1,206	201
Total Callander Bay watershed		2,574	430

4.1.2 Natural Runoff Loading

Natural phosphorus loads from the overland runoff in the watershed originate from phosphorus-bearing soils and decomposed organic matter. Groundwater may also contribute to natural phosphorus loads, but in Shield environments these contributions are most often negligible (Paterson et al., 2006) or occur in surficial soils as interflow and are included as surface runoff.

For Canadian Shield lakes, wetlands are the dominant factor controlling for phosphorus loads from the watershed (Dillon and Molot, 1997). Wetlands act as a sink for terrestrial organic matter from upstream sources and, under conditions of natural equilibrium, phosphorus is mineralized and much of the input exported downstream over an annual cycle (Devito and Dillon, 1993). The results of 20 years of monitoring data at 20 lake watersheds in central Ontario by the Dorset Environmental Science Centre (DESC) showed that natural phosphorus loads were observed to increase with wetland area following the equation (Paterson et al., 2006):

$$P \text{ export (kg/yr)} = (\text{catchment area (km}^2\text{)} * (0.47 * \% \text{ wetland area})) + 3.82$$

This equation was used to calculate the natural runoff of phosphorus in the watershed and loadings are summarized in Table 7 for individual subcatchment areas and Intake Protection Zone (IPZ) within the Callander vulnerable area for drinking water source protection. Catchment area in the equation above includes all natural undeveloped vegetated areas (all forest, wetland and grassland areas described in Section 2.2).



It should be noted that not all phosphorus loads to Wasi Lake are transported downstream via the lake’s outlet due to in-lake retention processes (uptake of dissolved phosphorus by plant life (algae) and settling of particulate phosphorus to lake sediments). Phosphorus loading rates in Table 7 do not account for phosphorus retention in Wasi Lake. The potential retention of phosphorus in Wasi Lake is discussed in Section 4.4.

The total phosphorus loading from natural runoff to Callander Bay is 3,871 kg/yr, 56% of which (2,162 kg/yr) originates from the vulnerable area of the Callander drinking water intake. This includes the contribution of 1,399 kg P/yr to Wasi Lake, a portion of which flows to Callander Bay via the Wasi River.

Table 7. Phosphorus Loading from Natural Runoff

Area		Forested Area	Wetland Area	Phosphorus Loading	Phosphorus Export
		(ha)	(%)	(kg/yr)	(kg/ha/yr)
By Subcatchments					
Wasi River Subwatershed	Wasi River catchment upstream of Wasi Lake	8,526	14	896	0.105
	Chiswick Creek catchment	1,676	16	193	0.115
	Total Wasi Lake and catchment (including Chiswick Creek and Wasi River catchments)	11,408	18	1,399	0.123
	Graham Creek catchment	5,110	24	779	0.152
	<i>Total Wasi River subwatershed</i>	<i>19,323</i>	<i>23</i>	<i>2,835</i>	<i>0.147</i>
La Vase River Subwatershed	Burford Creek catchment	1,119	42	264	0.236
	Tributary 1 catchment	245	44	60	0.247
	Tributary 2 catchment	174	45	43	0.249
	Cranberry Creek catchment	394	47	101	0.258
	<i>Total La Vase River subwatershed</i>	<i>1,931</i>	<i>44</i>	<i>469</i>	<i>0.243</i>
Bear-Boleau Creeks Subwatershed	Windsor Creek catchment	2,122	49	567	0.267
Total Callander Bay watershed		23,376	27	3,871	0.166
By Intake Protection Zone					
IPZ-1		10	38	2	0.217
IPZ-2		8	48	2	0.264
IPZ-3a		110	57	34	0.306
IPZ-3b		264	51	73	0.278
IPZ-3c		1,346	43	323	0.240
IPZ-3d		1,279	51	355	0.278
IPZ-3e		1,815	47	470	0.259
IPZ-3f		5,618	26	901	0.160
Total Intake Protection Zones		10,452		2,162	0.207



4.1.3 Internal Loading

Internal loading of phosphorus to a lake can occur either by release of phosphorus from anoxic lake sediments or through the decomposition and resultant release of phosphorus (mineralization) from organic sediments in shallower areas of warmer water bodies. Nearshore portions of Callander Bay are shallow and some sediment resuspension is therefore possible. There is a history of lumber industry in Callander and the deeper portions of the bay are reported to have accumulations of bark and sawdust. Their role in the phosphorus budget cannot be assessed from present knowledge. We note, however, that accumulations of organic material on the bottom of Precambrian Shield lakes is a common phenomenon and that wood bark is made up of larger particles of coarse woody debris. Large particles have less surface area per unit of mass and are therefore less reactive than the smaller particles of organic matter that typically accumulate in lake bottoms. As such, the oxygen dynamics of the bottom sediments are likely a more important factor than the nature of organic matter in determining internal loading.

Data collected by the NBMCA also indicate that periods of weak stratification do occur during the summer and temporary hypoxic (low oxygen) conditions develop in bottom waters. Both internal load mechanisms are therefore possible in Callander Bay. Internal loading is also likely in Wasi Lake. This lake is shallower than Callander Bay and periods of stratification-induced anoxia are unlikely to occur. Internal phosphorus load would therefore be most likely to occur due to slow mineralization processes or resuspension of sediments during wind mixing.

One way to assess whether internal phosphorus loading is an important contributor to phosphorus concentrations is through the use of lake modelling. In Section 4.4.2, a mass balance modelling technique is used to estimate total phosphorus concentrations in Callander Bay and Wasi Lake. The resulting modelled phosphorus concentrations compare very well with actual measured concentrations for Wasi Lake suggesting that all sources and processes have been accounted for in the model and that internal load contributions are not likely significant. For Callander Bay, the model overestimated phosphorus concentrations. If all other aspects of the model were accurate and internal phosphorus loading was significant, then the modelled values would be expected to underestimate concentrations because a loading source was not accounted for. The Callander Bay model, however, may need to be modified to account for influx and dilution by Lake Nipissing water during wind events (Section 4.4.2).

Based on comparison of modelled and measured phosphorus concentrations, internal phosphorus loading is not likely a significant component of the phosphorus budget for Wasi Lake. Its contribution to Callander Bay cannot, however, be ruled out without better knowledge of the mixing of Callander Bay with Lake Nipissing.

4.2 Anthropogenic Phosphorus Loading

Human sources of phosphorus include point and non-point sources. Point source loads are direct inputs from a specific pollution source such as a sewage treatment plant or an industrial effluent discharge, and can be determined directly from measurements of concentration and volume of these discharges. Non-point sources are diffuse sources, which include septic systems, urban runoff (storm water) or agricultural runoff. They are difficult to measure accurately on a site specific basis and so loadings are best estimated using export coefficients.



Point and non-point sources may discharge directly to the water body or may enter via runoff from the watershed upstream of the water body.

4.2.1 Septic Systems

Calculation of phosphorus loads from septic systems follows the approach recommended by MOE's Ontario's Lakeshore Capacity Model (Paterson et al., 2006) where:

$$\text{Load per septic system (kg)} = \text{per capita phosphorus load (kg/capita/yr)} * \text{occupancy rate (capita yrs/yr)}$$

Based on a review of measured data and the literature pertaining to phosphorus concentrations in septic systems and average water usage, the per capita phosphorus load from septic systems is estimated to be 0.66 kg/capita/yr, 0.44 kg/capita/yr and 0.22 kg/capita/yr for septic systems within 100 m, between 100 and 200 m, and between 200 and 300 m of the shoreline, respectively (Hutchinson 2002, Paterson et al. 2006). The majority of shoreline residences in the Callander Bay watershed are likely occupied year-round, therefore the permanent occupancy rate of 2.56 capita yrs/yr was chosen for all septic systems. The total number of septic systems was estimated by the NBMCA as the number of un-serviced lots that lie (wholly or in part) within 100m, between 100 to 200m and between 200 to 300m of the shoreline of Callander Bay or water courses and water bodies draining to Callander Bay.

While shoreline septic systems can be a significant source of phosphorus to lakes, recent scientific studies have shown that much of the septic phosphorus load is attenuated by acidic and mineral-rich soils found in the Precambrian Shield. Mechanistic evidence (Stumm and Morgan, 1970; Jenkins *et al.*, 1971; Isenbeck-Schroter *et al.*, 1993) and direct observations made in septic systems (Willman *et al.*, 1981; Zanini *et al.*, 1997; Robertson *et al.*, 1998; Robertson, 2003) all show strong adsorption of phosphate on charged soil surfaces and reactions to form minerals with iron (Fe) and aluminum (Al) in soil. These mineralization reactions, in particular, appear to be favoured in acidic and mineral rich groundwater in Precambrian Shield settings (Robertson *et al.*, 1998; Robertson, 2003), such that over 90% of septic phosphorus may be immobilized. The mineralization reactions appear to be permanent (Isenbeck-Schroter *et al.*, 1993). Recent studies conclude that most septic phosphorus may be stable within 0.5 m of the tile drains in a septic field (Robertson *et al.*, 1998, Robertson, 2003).

Trophic status modelling also supports the mechanistic and geochemical evidence of phosphorus attenuation by soils. Dillon *et al.* (1994), for example, reported that only 26% of the potential loading of phosphorus from septic systems around Harp Lake, Muskoka, could be accounted for in the measured phosphorus budget of the lake. The authors attributed the variance between measured and modelled estimates of phosphorus to retention of septic phosphorus in thick tills in the catchment of Harp Lake.

Given strong scientific evidence supporting attenuation of septic phosphorus by soils, combined with the large areas of deep till deposits in the watershed of Callander Bay, it is likely that not all phosphorus from shoreline septic systems is delivered to Callander Bay. Much of the phosphorus load from septic systems in areas of deep soils that are rich in iron and aluminum is likely attenuated with only a fraction of the potential load reaching the lake. Potential septic system phosphorus loads to Callander Bay were therefore estimated assuming 1) that all phosphorus reaches the bay (0% attenuation by soils), and 2) that only 26% of the septic



phosphorus moves to the bay (74% attenuation by soils), in line with the findings of Dillon *et al.* (1994) (Table 8). Additional site specific information regarding soil conditions of septic systems is required to better estimate the actual phosphorus loads to Callander Bay from this source.

By definition, the vulnerable area (area of all IPZs) of the Callander drinking water intake for Drinking Water Source Protection includes all water bodies contributing water to Callander Bay and land area draining to those water bodies to a maximum setback of 120 m from the high water mark. As such, nearly all shoreline septic system phosphorus loads to Callander Bay originate from the vulnerable area of the Callander drinking water intake.

Septic systems contribute as much as 1,287 kg of phosphorus per year to Callander Bay and 345 kg per year to Wasi Lake assuming that 100% of the phosphorus is transported from the septic system to the adjacent water course or lake. Phosphorus loading from septic systems may be as little as 335 kg/yr and 90 kg/yr for Callander Bay and Wasi Lake, respectively if soils are assumed to attenuate 74% of the phosphorus. Additional site specific information regarding septic systems, their locations (distance from the shoreline) and soil conditions is required to more accurately estimate septic system loading.



Table 8. Phosphorus Loading from Septic Systems

Area		Distance from Shoreline				Phosphorus Load (kg/yr)	
		<= 100 m	>100 - 200 m	>200 - 300 m	Total	0% Attenuation	74% Attenuation
By Subcatchment							
Wasi River Subwatershed	Wasi River catchment upstream of Wasi Lake	26	21	29	76	84	22
	Chiswick Creek catchment	25	9	8	42	57	15
	<i>Total Wasi Lake and catchment (including Chiswick Creek and Wasi River catchments)</i>	170	31	40	241	345	90
	Graham Creek catchment	29	43	19	91	108	28
	<i>Total Wasi River subwatershed</i>	278	152	94	524	694	180
La Vase River Subwatershed	<i>Total La Vase River subwatershed</i>	66	68	45	179	213	55
Bear-Boleau Creeks Subwatershed	Windsor Creek catchment	59	75	36	170	204	53
Callander Bay shoreline		93	13	7	113	176	46
Total Callander Bay watershed		496	308	182	986	1,287	335
By Intake Protection Zone¹							
IPZ-1		37	0		37	63	16
IPZ-2		3	0		3	5	1
IPZ-3a		59	7		66	108	28
IPZ-3b		1	0		1	2	0
IPZ-3c		94	27		121	189	49
IPZ-3d		61	17		78	122	32
IPZ-3e		66	28		94	143	37
IPZ-3f		175	14		189	311	81
Total Intake Protection Zones		496	93		589	943	245

Notes: ¹By definition, the IPZs include land area of up to 120 m from the high water mark of a water body or water course. Some septic systems in the IPZs may lie beyond 100 m and therefore contribute a lower phosphorus load than reported.

4.2.2 Agriculture

Specific agricultural practices in the Callander watershed are not specifically quantified but include livestock operations (horse, sheep, cattle), pasture and some cropland. In the absence of specific data, a general export coefficient of 0.30 kg/ha/year was chosen for agricultural areas, as the mean export from 198 watersheds draining cropland in North America calculated by Chambers and Dale (1997; range = 0.12-0.39 kg/ha/yr) and recommended for use in MOE's Lakeshore Capacity Model (Paterson et al., 2006). This coefficient is higher than the accepted export from cleared land/pasture (0.098 kg/ha/yr; Paterson et al., 2006), but recognizing that there are several livestock operations (with manure piles, feedlots) and some cropland with higher phosphorus exports, phosphorus supply from agricultural lands in the Callander Bay watershed is likely higher than from pasture alone. It should also be noted that parkland and golf courses were included in the classification of agricultural land use and that these would also have different phosphorus exports depending on fertilizer use, etc. More detailed agricultural land use data (e.g., area and type of cropland, number and type of livestock) and differentiation



of parkland lawns and golf courses are required to refine the following estimate of phosphorus loading from agricultural lands to Callander Bay.

Phosphorus loading from agricultural lands is provided in Table 9 by subcatchment area and by Intake Protection Zone (IPZ) of the Callander vulnerable area for drinking water source protection.

The total phosphorus loading from agricultural runoff to Callander Bay is 984 kg/yr, 47% of which (436 kg/yr) originates from the vulnerable area of the Callander drinking water intake. Agricultural runoff contributes a total of 348 kg P/yr to Wasi Lake.

4.2.3 Urban Runoff

Urban runoff includes runoff from paved areas, disturbed surfaces, parking lots, urban lawns (fertilized and non-fertilized) and rooftops. This runoff can contain phosphorus from direct additions (i.e., fertilizers, urban dust, animal droppings) and indirect sources such as erosion induced by increased runoff. The characteristics of urban runoff will therefore vary with the contributing areas and sources.

For the Callander Bay watershed, an export coefficient of 1.32 kg/ha/yr for developed urban areas was chosen from the nutrient model of Winter *et al.* (2003) for Lake Simcoe and recommended by the MOE for mid to high density urban areas. The resultant phosphorus loading from urban runoff are summarized in Table 9 by subcatchment area and Intake Protection Zones (IPZ) of the Callander vulnerable area for drinking water source protection.

The total phosphorus loading from urban runoff to Callander Bay is 512 kg/yr, 71% of which (296 kg/yr) originates from the vulnerable area of the Callander drinking water intake. Urban runoff contributes a total of 32 kg P/yr to Wasi Lake.



Table 9. Phosphorus Loading from Agricultural and Urban Lands

Area		Agricultural Land Area	Agricultural Phosphorus Loading	Urban Land Area	Urban Phosphorus Loading
		(ha)	(kg/yr)	(ha)	(kg/yr)
By Subcatchments					
Wasi River Subwatershed	Wasi River catchment upstream of Wasi Lake	695	208	11	14
	Chiswick Creek catchment	187	56	3	3
	<i>Total Wasi Lake (including Chiswick Creek and Wasi River catchments)</i>	<i>1,160</i>	<i>348</i>	<i>24</i>	<i>32</i>
	Graham Creek catchment	1,436	431	58	77
	<i>Total Wasi River subwatershed</i>	<i>2,829</i>	<i>849</i>	<i>161</i>	<i>213</i>
La Vase River Subwatershed	Burford Creek	71	21	41	54
	Tributary 1	42	13	39	52
	Tributary 2	30	9	40	53
	Cranberry Creek	63	19	17	23
	<i>Total La Vase River subwatershed</i>	<i>205</i>	<i>61</i>	<i>138</i>	<i>182</i>
Bear-Boleau Creeks Subwatershed	Windsor Creek	247	74	89	117
Total Callander Bay watershed		3,280	984	388	512
By Intake Protection Zone					
IPZ-1		7	2	10	13
IPZ-2		4	1	7	10
IPZ-3a		24	7	28	36
IPZ-3b		21	6	7	9
IPZ-3c		117	35	76	101
IPZ-3d		139	42	44	58
IPZ-3e		549	165	42	56
IPZ-3f		593	178	11	14
Total Intake Protection Zones		1,455	436	224	296

4.2.4 Urban Waterfowl

Callander Bay has a large littoral area and abundant wetlands that provide exceptional natural habitat for waterfowl. There is concern, however, about large numbers of nuisance waterfowl (Canada geese) near waterfront areas of the Town, for example, at Centennial Park. In urban areas, habitat for waterfowl is increased due to human alteration of the landscape, often resulting in large numbers of waterfowl. Nuisance waterfowl sites in urban settings are typically associated with areas where forage materials are abundant (e.g., short grass) and where there is no emergent shoreline vegetation such that foraging birds can easily seek refuge on water if disturbed. Golf courses, parks, beaches and large expanses of lawns with unimpeded access to water often attract foraging waterfowl in large numbers, which can contribute significantly to the phosphorus load in urban settings.



The following provides an estimate of potential phosphorus loading from geese at Centennial Park based on observations of goose numbers in the summer and fall of 2010 and published estimates of phosphorus contributions from goose feces.

Following the methods of Moore et al. (1998), the average adult Canada goose (*Branta canadensis maxima*) exports approximately 1.22 g of phosphorus per day in feces. Given that geese do not reside at the park for the full day, not all of this phosphorus is input directly to the bay adjacent to the park, some is likely to be exported to other areas removed from the water or even other watersheds not draining to the bay. For this exercise, therefore, we assume that 50% of the potential phosphorus from the goose droppings (0.66 g) is delivered to the bay in the vicinity of Centennial Park. The occurrence of geese at Centennial Park is variable, with numbers ranging from approximately 25 to 70 geese per day beginning at the end of July to the end of September and with migratory flocks of over 100 birds in October (based on observations by J. Celentano, 2010). Assuming an average of 50 birds per day in August and September and 75 birds per day in October, approximately 3.0 kg of phosphorus is exported to the Bay from geese residing at Centennial Park. This is a gross estimate based on casual counts of birds at Centennial Park. There have been other reported cases of large numbers of foraging geese at Osprey Links golf course and at the ball field west of Callander Bay Drive. More detailed observational data would be required to more accurately determine the number of geese residing and foraging in urban areas near Callander Bay and to calculate the potential phosphorus load from this source.

While the export of phosphorus from urban waterfowl at Centennial Park is small in comparison to other sources, the export occurs over a short time period in a concentrated area. This source of phosphorus could therefore be a significant contributor to the phosphorus load during late summer and result in localized algal bloom activity in the beach area of the park. Moreover, the phosphorus-rich goose feces may be directly deposited to the lake and or washed into the lake from nearshore areas contributing to phosphorus enrichment of the sediments, which can lead to enhanced growth of nuisance aquatic macrophytes.

Although geese in the Callander Bay watershed are not confined to the waterfront areas of the Bay, they were considered explicitly for these areas because of the direct access to open water and their known areas of congregation in urban areas. Waterfowl also use wetlands and open fields as habitat, but loading from these areas would be incorporated in the respective export coefficients for phosphorus runoff. Any additional known areas of congregation (i.e., Wasi Lake) should be identified and incorporated into future budgets.

Canada geese residing at Centennial Park contribute approximately 3.0 kg of phosphorus per year to Callander Bay based on estimates of bird numbers and resident days.

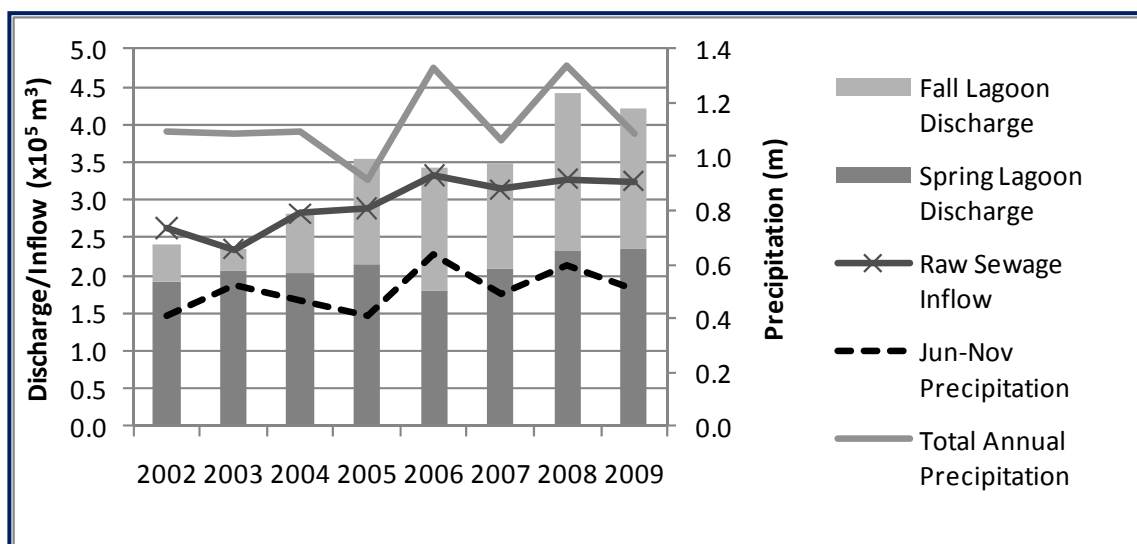
4.2.5 Point Sources

The Callander Waste Water Treatment Plant (WWTP), servicing approximately 1,400 people, is the only known point source of phosphorus to Callander Bay. The treatment system is a Class 1 Treatment system that consists of two seasonal release waste stabilization lagoons with a total volume of 264,000 m³, and a 4,600 m³ sludge disposal lagoon. The lagoons are treated with ferric sulphate to remove phosphorus and are released in spring and fall to Cranberry Creek, which then discharges to Callander Bay.



The total discharge of treated effluent to Callander Bay from the WWTP has increased steadily since 2002 from 240,714 m³ to 422,354 m³ in 2009 with the increase occurring during the fall lagoon release period (Figure 15). The reason for the higher volume of effluent discharged during the fall is not known conclusively. Raw sewage flows to the lagoons have increased over the same time period, but there is no clear relationship between precipitation patterns and sewage inflow or discharge volumes (Figure 15).

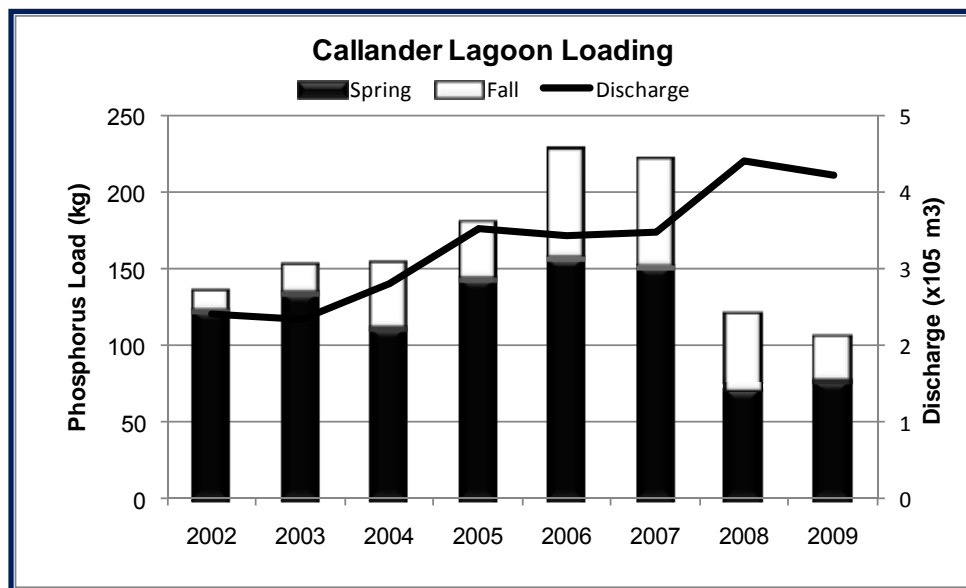
Figure 15. Annual raw sewage inflow and effluent discharge volumes at the Callander WWTP and precipitation (2002-2009).



Phosphorus loads to Callander Bay increased proportionally with discharge until 2008, when improvements were implemented at the WWTP that effectively lowered the concentration of phosphorus in the treated effluent resulting in significantly lower phosphorus loads (Figure 16). The mean total phosphorus loading in 2008 and 2009 was 113.48 kg/yr in comparison to the 2002-2007 average loading of 178.72 kg/yr.

The Callander WWTP contributes a loading of 113 kg P/yr to Callander Bay, which reflects current phosphorus removal abilities at the plant.

Figure 16. Phosphorus loads and effluent discharge volumes to Callander Bay from the Callander WWTP(2002-2009).



4.3 Total Loading

The total phosphorus loading to Callander Bay from all sources described above ranges from 6,906 kg/yr assuming all septic phosphorus migrates to the bay, to 5,953 kg/yr assuming that soils attenuate 74% of the septic load. These loading estimates, however, do not take into account the potential for phosphorus retention in Wasi Lake. Not all phosphorus contained in a lake is passed on to downstream lakes because a portion of the phosphorus is lost from the water column to the sediments.

The amount of phosphorus retained in a lake (R) is a function of the relationship between its areal water load (q) and the settling velocity (v) of phosphorus, where:

$$R = v/(v+q)$$

The settling velocity (v) of phosphorus is estimated to be 12.4 m/yr for stratified oligotrophic lakes on the Precambrian Shield and 7.2 m/yr for those stratified lakes with anoxic hypolimnia (Dillon *et al.* 2006). These settling velocities, however, are not applicable to Wasi Lake. Wasi Lake is very productive (eutrophic) and because it is shallow (mean depth = ~2 m), its water column is easily mixed by wind which prevents thermal stratification and maintains oxic conditions near the sediments. In shallow lakes, wind mixing results in slower settling of phosphorus (and reduced in-lake retention) in comparison to stratified lakes with similar areal water loads. There are no known published estimates of settling velocities for shallow, productive lakes like Wasi Lake. Based on comparisons between measured and modelled phosphorus concentrations in shallow lakes in the City of Elliot Lake (HESL, 2010) and in Seguin Township (AECOM, 2009), a settling velocity of 3.6 m/yr was found to best approximate phosphorus retention in shallow lakes. Using this settling velocity, the phosphorus retention (R)



in Wasi Lake is 0.29; that is 29% of the phosphorus in Wasi Lake is lost to the sediments and not passed downstream to Callander Bay. This retention is consistent with the results of detailed phosphorus mass balances constructed by measuring all inputs and losses of phosphorus to Rice Lake and Sturgeon Lake by the Ontario Ministry of the Environment (Hutchinson et al., 1994). The average phosphorus retention for three years was 21% for Rice Lake (average depth = 2.4 m) and 24% for Sturgeon Lake (average depth = 3.5 m). Wasi Lake is most similar to Rice Lake with respect to average depth and so the retention of 21% was chosen for the model.

A summary of phosphorus loading to Callander Bay is presented in Table 10 by subcatchment area and by Intake Protection Zones (IPZs) of the Callander drinking water vulnerable area for drinking water source protection considering a 21% retention of phosphorus in Wasi Lake. The relative contribution of the individual sources to the total loading to Callander Bay and Wasi Lake is summarized in Figure 17 assuming no attenuation of septic phosphorus by soils.

The total phosphorus loading to Callander Bay from all sources is 6,426 kg/yr assuming that all septic phosphorus reaches the bay and 5,527 kg/yr assuming 7% retention of septic phosphorus by soils and considering phosphorus retention of 21% in Wasi Lake. Between 51% and 53% of the total load to Callander Bay is supplied from land area encompassed by the Callander drinking water IPZs. For Wasi Lake, the total phosphorus loading from all sources is 2,283 kg/yr or 2,027 kg/yr assuming 0% and 7% attenuation of septic phosphorus by soils, respectively.

Assuming that all septic phosphorus reaches the lake, 54% of the total loading to Callander Bay is from natural sources. Septic systems are the largest human sources of phosphorus, supplying ~19% of the total loading to Callander Bay, followed by agriculture (14%) and urban runoff (8%). Only 2% of the phosphorus loading is supplied by the Callander Waste Water Treatment Plant. A large portion (up to 59%) of the loading from human sources to Callander Bay originates from the IPZs.

Assuming that all septic phosphorus reaches the lake, 68% of the total loading to Wasi Lake is from natural sources. Septic systems and agriculture contribute nearly equal loading (~15%) and urban runoff contributes only 2% of the total loading to Wasi Lake.



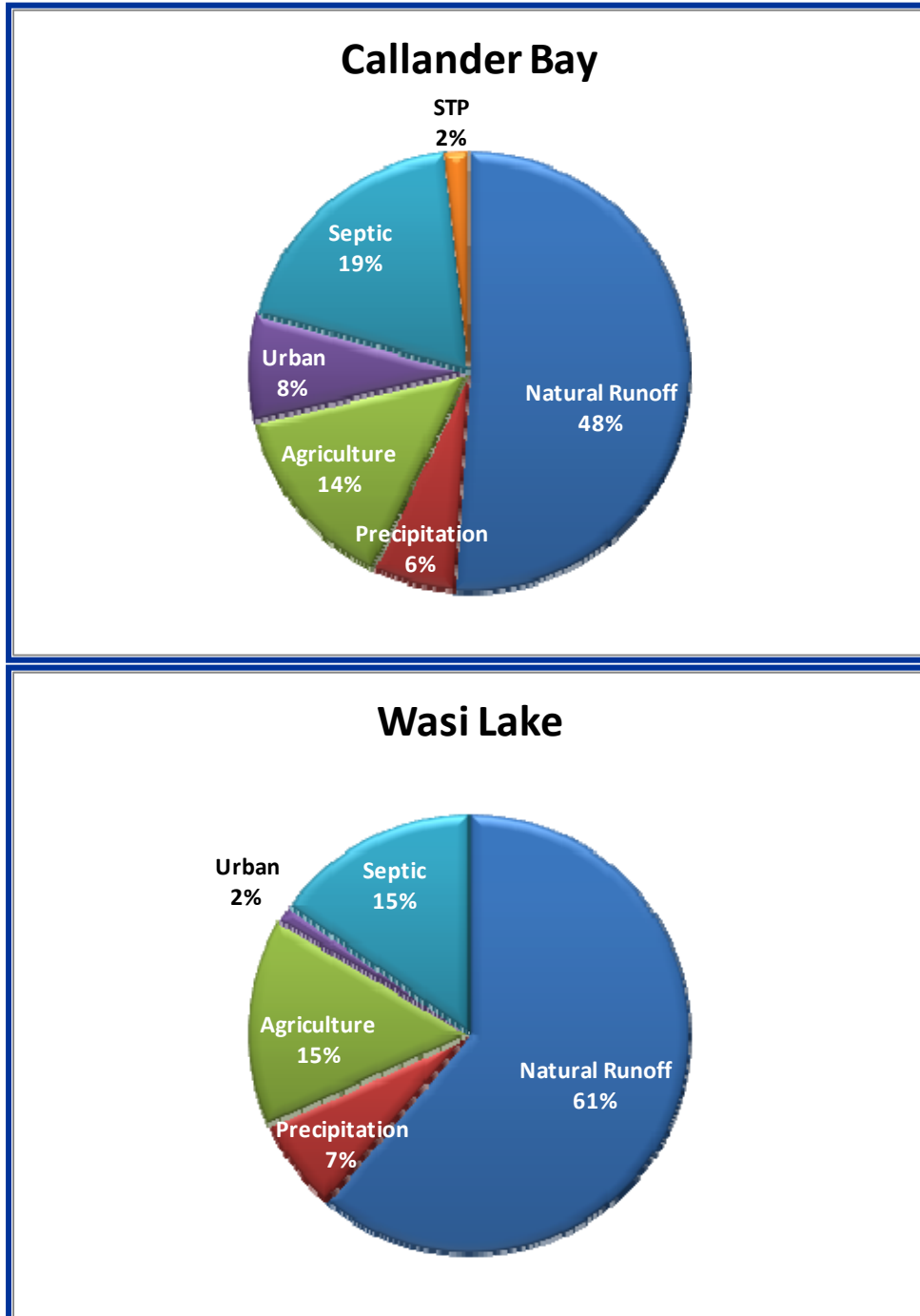
Table 10. Total Phosphorus Loading Summary for Callander Bay and Watershed

Area	Loading (kg/yr)								Total Loading (kg/yr)	% Load to Callander Bay	Total Loading (kg/yr)	% Load to Callander Bay
	Natural Runoff	Precipitation	Agricultural Runoff	Urban Runoff	Septic		STP	Urban Water-fowl				
					0% Attenuation	74% Attenuation						
	0% Attenuation		74% Attenuation									
By Subcatchments												
Wasi River catchment upstream of Wasi Lake	896	28	208	14	84	22			1,230	19	1,168	21
Chiswick Creek catchment	193	9	56	3	57	15			318	5	276	5
<i>Total Wasi Lake and catchment</i>	<i>1,399</i>	<i>159</i>	<i>348</i>	<i>32</i>	<i>345</i>	<i>90</i>			<i>2,283</i>	<i>36</i>	<i>2,027</i>	<i>37</i>
Graham Creek catchment	779	19	431	77	108	28			1,413	22	1,333	24
<i>Total Wasi River sub watershed</i>	<i>2,541</i>	<i>187</i>	<i>775</i>	<i>206</i>	<i>694</i>	<i>180</i>			<i>4,404</i>	<i>69</i>	<i>3,890</i>	<i>70</i>
Burford Creek catchment	264	11	21	54					349	5	349	6
Tributary 1 catchment	60	3	13	52					128	2	128	2
Tributary 2 catchment	43	1	9	53					107	2	107	2
Cranberry Creek catchment	101	6	19	23					149	2	149	3
<i>Total from La Vase River sub watershed</i>	<i>469</i>	<i>21</i>	<i>61</i>	<i>182</i>	<i>213</i>	<i>55</i>			<i>947</i>	<i>15</i>	<i>789</i>	<i>14</i>
Windsor Creek catchment	567	20	74	117	204	53			983	15	831	15
Callander Bay		201										
Total Callander Bay and watershed¹	3,283	397	911	505	1,215	316	113	3	6,426	100	5,527	100
By Intake Protection Zone (IPZ)												
IPZ-1	2		2	13	63	16		3	83	1	36	1
IPZ-2	2		1	10	5	1			18	0	15	0
IPZ-3a	34		7	36	108	28			185	3	105	2
IPZ-3b	73		6	9	2	0			90	1	89	2
IPZ-3c	323		35	101	189	49			649	10	508	9
IPZ-3d	355		42	58	122	32			577	9	487	9
IPZ-3e	470		165	56	143	37			834	13	728	13
IPZ-3f	901		178	14	311	81			1,405	22	1,174	21
Total IPZs	1,900		385	292	852	222			3,429	53	2,798	51

Notes: ¹values adjusted to account for a 21% retention of phosphorus in Wasi Lake; septic systems around the shoreline of Callander Bay are included in the totals for Total Callander Bay and Watershed area only (not in individual subwatershed areas)



Figure 17. Relative contribution of phosphorus sources to the total phosphorus loading to Callander Bay and Wasi Lake assuming no attenuation of septic phosphorus by soils³.



³ Natural load = Natural runoff + Precipitation to open water

4.4 Phosphorus Budget Validation

4.4.1 Measured Versus Modelled Loading from the Wasi River

On average, the Wasi River supplies a loading of 4,105 kg/yr from all sources in its watershed to Callander Bay based on measured long-term total phosphorus concentration data (2004-2010) and limited hydrological data (2009) (Table 11), which represents an export of 0.184 kg/ha/yr from the Wasi River watershed to Callander Bay. More than 55% of the loading occurs during the spring (March to May) when runoff from the watershed is greatest.

The measured loading from Wasi River differs by 23% from the loading of 4,834 kg/yr that was calculated using an export coefficient approach (modelled loading) and is greatly improved to a difference of only 4 to 16% when phosphorus retention in Wasi Lake is accounted for in the budget (modelled loading = 3,924-3,465 kg/yr) (Table 12). The close agreement between measured and modelled phosphorus loading from the Wasi River provides confidence in the chosen export coefficients to estimate phosphorus load from different land use types in the Callander Bay watershed as well as the estimation of phosphorus retention in Wasi Lake. Furthermore, the closer agreement between measured and modelled loading when no septic system phosphorus is assumed to be attenuated by soils suggests that septic system phosphorus may be mobile in the Wasi River watershed.

Table 11. Measured Phosphorus Loading from the Wasi River to Callander Bay

Month	Mean Total Phosphorus Concentration ($\mu\text{g/L}$)	2009 Depth of Runoff (m)	Total Phosphorus Load (kg)	% of Annual Phosphorus Load
Jan	35.9	0.036	307	7.47
Feb	35.9	0.021	174	4.24
Mar	35.9	0.054	460	11.20
Apr	44.2	0.119	1,235	30.10
May	42.7	0.059	594	14.47
Jun	50.2	0.021	252	6.13
Jul	53.2	0.017	217	5.28
Aug	43.8	0.011	111	2.71
Sep	37.8	0.006	51	1.23
Oct	27.6	0.019	123	3.00
Nov	35.9	0.041	343	8.35
Dec	35.9	0.028	239	5.83
Annual	39.9	0.432	4,105	100.00



Table 12. Comparison of Measured and Modelled Loading from the Wasi River to Callander Bay Assuming 21% Retention of Phosphorus in Wasi Lake

Scenario	Loading (kg/yr)
Measured	4,105
Modelled with no septic attenuation	3,924
<i>% difference from measured</i>	4
Modelled with 74% septic attenuation	3,465
<i>% difference from measured</i>	16

4.4.2 Phosphorus Concentration Modelling

Wasi Lake

The modelled total phosphorus concentration in Wasi Lake ranges from 32.4 µg/L assuming that all of the phosphorus from shoreline septic systems reach the lake, to 28.8 µg/L assuming that 74% of the septic phosphorus is attenuated by soils.

The modelled phosphorus concentrations reflect long-term steady state conditions in Wasi Lake and compare well with measured phosphorus concentrations observed in 2007 and 2009 (Table 13). In 2008, mean measured phosphorus concentration was elevated by more than 35% over the 2007/2009 average, which exceeds the inter-annual variability of ~20% that is typically seen in Shield lakes monitored by MOE’s Lake Partner Program (Bev Clark, p. comm.). This suggests that phosphorus concentrations were anomalously high in 2008, likely due to the high precipitation that occurred that summer (Figure 3), and are not likely representative of long-term mean conditions in the lake⁴.

Both modelled phosphorus concentrations (with and without septic phosphorus attenuation by soils) are within 20% of the mean measured concentration in 2007 and 2009 (Table 13), which is an acceptable degree of error (Paterson et al., 2006) and provides confidence the estimated phosphorus loads to Wasi Lake. There is a better agreement between measured and modelled total phosphorus concentrations in Wasi Lake when no attenuation of septic phosphorus is considered in the model. This suggests that phosphorus from shoreline septic systems is mobile and reaches the lake.

⁴ Reviewers suggested that removal of an illegal dam at the outflow of Wasi Lake in the autumn of 2007 by MNR may have influenced phosphorus concentrations in the lake in 2008. Although dam removal would have reduced lake levels we note that a) concentrations in 2009 were reduced again and b) that precipitation was 27% higher than average in 2008, suggesting that the dam removal did not alter lake nutrient levels.



Table 13. Comparison between Measured and Modelled Phosphorus Concentration in Wasi Lake

Scenario		Total Phosphorus Concentration	
		(µg/L)	No. of Samples
Measured	2007	33.7	14
	2008	44.8	16
	2009	32.9	22
	Mean	37.1	
	Mean 2007 and 2009	33.3	
Modelled	No attenuation of septic P	32.4	
	<i>% difference from Mean (2007/2009)</i>	-2.7	
	74% attenuation of septic P	28.8	
	<i>% difference from Mean (2007/2009)</i>	-13.5	

Callander Bay

The modelled total phosphorus concentration in Callander Bay ranges from 25.4 µg/L assuming that all of the phosphorus from shoreline septic systems reach the lake, to 22.8 µg/L assuming that 74% of the septic phosphorus is attenuated by soils. These estimates differ by 18% and 6%, respectively from the observed mean ice free phosphorus concentration in the bay (2007-2009) of 21.5 µg/L.

Both modelled phosphorus concentrations (with and without septic phosphorus attenuation by soils) are within 20% of the mean measured concentration, which is an acceptable degree of error (Paterson et al., 2006) and provides confidence the estimated phosphorus loads to Callander Bay. There is a better agreement between measured and modelled total phosphorus concentrations in Callander Bay when attenuation of septic phosphorus is considered in the model. This suggests that mobility of phosphorus from shoreline septic systems may be reduced by soils.

While results of the modelled phosphorus concentrations are acceptable, we note that there is likely error in the modelled values due to:

1. Inability of the model to account for mixing with Lake Nipissing. As the phosphorus concentrations in Lake Nipissing are lower than those in Callander Bay, inflow of water from Lake Nipissing would result in a reduction of the phosphorus concentration in Callander Bay given the same loading estimates,
2. Inability to accurately determine settling velocity and the loss of phosphorus to the sediments. A settling velocity of 12.4 m/yr was assumed for the above model predictions. This settling velocity is greater than that used for Wasi Lake (3.6 m/yr) as Callander Bay is deeper and does undergo periodic stratification.
3. Over-estimation of total phosphorus loading from natural sources to Callander Bay. Wetland areas represent 36% of the natural area (i.e., forest, wetland and grassland) in the Callander Bay watershed excluding the Wasi Lake watershed. This high percentage of wetland exceeds the range of % wetland in the catchment of lakes used to define the wetland equation (Paterson et al., 2006) and therefore the equation may not be best



suited for use in this case. While an over estimation of the phosphorus loading to the bay is possible, we note that there was excellent agreement between measured and calculated loads in the Wasi River, which encompasses a large portion of the total watershed area and contributes 65% of the total load to the bay.

4.4.3 Validation Summary

Overall, there is a high degree of confidence in the total estimated phosphorus loading to Callander Bay from the Wasi River. This is supported by the agreement between measured and modelled loadings for the Wasi River (Section 4.4.1) and measured and modelled phosphorus concentrations in Wasi Lake (Section 4.4.2).

Results of the validation also suggest that phosphorus from septic systems is mostly mobile in the Wasi River subwatershed. While there are deep, mineral-rich soils in much of the Wasi River subwatershed that would potentially prevent or slow the movement of phosphorus to Callander Bay, hydrological conditions such as an elevated water table that reaches the septic beds may be preventing the soil attenuation processes. Site specific observations of shoreline septic systems that document soil type and depth, distance from the shoreline and depth of the water table are required to more confidently confirm the degree of septic phosphorus mobility.

For Callander Bay, while measured and modelled total phosphorus concentrations are comparable (within 20%), some error is expected due to a combination of inaccuracies in model inputs (i.e., settling velocity) and the inability of the model to account for mixing with Lake Nipissing, but also from an overestimation of the total phosphorus loading from natural areas to the bay. Based on the excellent validation results from the Wasi River and Wasi Lake modelling, however, we suspect that the loading estimates for Callander Bay are likely a good approximation of the actual loads and that inaccuracies in the model result from the inability to accurately address phosphorus retention and mixing with Lake Nipissing.

5. Information Gaps and Future Monitoring Requirements

It should be noted that while total loadings from the Wasi River subwatershed to Callander Bay appear to be accurate, there remain information gaps that lead to uncertainty in the relative contribution of different sources of phosphorus to Callander Bay, as well as the response of Callander Bay, particularly for agricultural activities.

Recommendation 1. Refinement of Agricultural and Urban Estimates

Runoff from agriculture and urban areas made up ~20% of the total load to Callander Bay and ~50% of the anthropogenic load and so these areas represent sources that could potentially be reduced. Further detail is needed; however, in order to a) confirm the estimates and b) identify those areas for mitigation.



The agricultural land cover class does not differentiate between agricultural activities or identify livestock operations that would supply significantly different amounts of phosphorus.

1.1 - We therefore recommend that agricultural land use classifications be refined to include areas devoted to pasture, field crops and row crops and to separate out golf courses and manicured lawns from agricultural practices. This could be done by targeted aerial photography or more detailed analysis of satellite data.

1.2 - We recommend that information regarding the number and type of animals on watershed farms, the location and sizes of feed lots and manure piles be collected to a) aid in the determination of phosphorus loading from livestock operations and b) identify the best options for BMP implementation. This could be most accurately done by a direct watershed inventory of each agricultural operation.

Recommendation 2. Continue and Expand Tributary Monitoring

The availability of measured phosphorus concentration data was very useful and showed that the loading estimates for the Wasi River were valid. Our review of hydrologic data showed, however, that flow was variable between years and so loading would also be expected to be. Callander Bay has a hydrologic flushing rate of ~1.9 years (independent of wind derived mixing of water from Lake Nipissing) and so would be expected to reflect the influence of the previous two years of watershed loading in any one year.

2.1 - We therefore recommend that the Wasi River monitoring conducted by the NBMCA continue for at least two more years in order to assess interannual variability in loading to Callander Bay. Sampling should continue year round with particular focus on the spring freshet period (late March to end of May) in future years. Although more detailed data review may allow some of the existing sites to be dropped for further surveys, we note that the costs of sampling are largely associated with labour, and that additional phosphorus analyses are relatively inexpensive.

2.2- At the time of report production, phosphorus concentration data were only available for June to September, 2009 for Chiswick and Graham Creeks. Additional data are required to estimate concentrations for the remainder of the year particularly in spring (April, May), and to establish mean annual concentrations and to refine loading estimates. These data could then be used to validate the export coefficient loading estimates at discrete points along the tributaries as was done for Wasi River.

2.3 - Additional monitoring sites should be established in agricultural streams to collect phosphorus concentration and flow data. This information would allow calculation of land use-specific phosphorus export from agricultural lands and be used to inform management opportunities. The number and locations of additional sites and types of agricultural streams should be determined based on the results of the watershed inventory of different agricultural practices.

2.4 – Loading estimates require accurate assessment of flow as well as concentration. Although the WSC gauge near Astorville provided useful estimates of flow for the other sites on the Wasi River, the accuracy of prorating varies and decreases with smaller streams. We therefore recommend that flow measurements be taken at the same time



as water quality samples at a) water quality sites on Chisholm and Graham Creeks and b) any agricultural streams identified for sampling in Recommendation 2.3.

Recommendation 3. Callander Bay Phosphorus Load

Internal loading of phosphorus from anoxic sediments in Callander Bay is a potential source that could be neither confirmed nor excluded in the present study. Internal phosphorus can represent a significant source to some cyanobacteria species that can alter their buoyancy to take advantage of it.

3.1 – We recommend that 4-6 profiles of phosphorus, temperature and dissolved oxygen be taken at 1-m intervals from two deep locations in Callander Bay in August and September. Half the profiles should be taken after and during periods of calm conditions and half when wind has altered stratification.

The phosphorus budget and preliminary model substantially overestimated phosphorus concentrations in Callander Bay but provided good agreement for Wasi Lake. The export coefficient model agreed well with measured loads in the Wasi River. The phosphorus budget therefore appears to be reliable but the model for the response of Callander Bay is not. Northland Engineering (1993) reported that water from Callander Bay does mix with Lake Nipissing. This influence could be substantial but is un-quantified.

3.2 - We recommend that further investigations of the mixing of Lake Nipissing with Callander Bay be investigated, to assess the degree of mixing and resultant effect on phosphorus concentrations in the Bay.

Recommendation 4. Septic Systems

Refinement of initial estimates by the NBMCA showed that there are 986 septic systems within 300m of Callander Bay and its tributary streams, and that 589 of these lie within the IPZs that were identified for Source Water Protection Planning. The significance of septic systems as a phosphorus source is uncertain because of uncertainty in the retention of phosphorus in soils, although failing systems will certainly represent a phosphorus source.

4.1 – Periodic septic inspections are therefore recommended for systems within 100m of an IPZ. Any systems identified as failing should be replaced or remediated.

Septic systems contribute anywhere from ~6% to ~20% of the total loading of phosphorus to Callander Bay, dependent on assumptions regarding phosphorus mobility from that source. The study suggests that septic system phosphorus is mobile, based on agreement with modelled and measured estimates for Wasi Lake. The importance of soils in phosphorus mobility, however, suggests that some confirmation of soil characteristics would inform the assumptions.

4.2 - We recommend that samples of B horizon soils be taken from 6 locations within 100m of surface water in each of the Callander Bay and Wasi Lake watersheds and analysed for mineral content and phosphorus adsorption capability to inform the likelihood of septic phosphorus mobility.

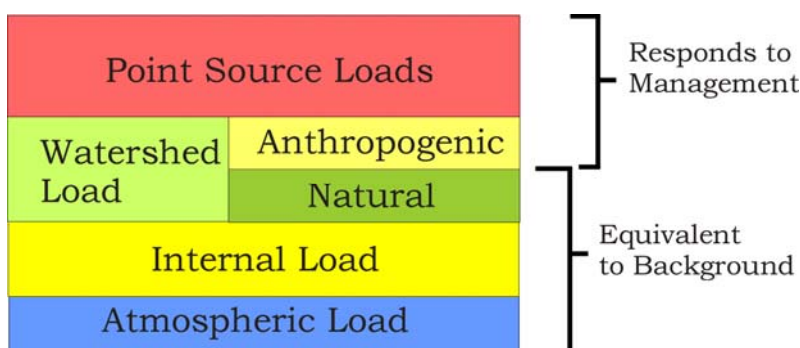


4.3 – We recommend that mapping of surficial soil depths be obtained for the watershed to identify areas of high risk from septic systems based on soil depth, slope and proximity to surface waters. These areas would be targeted for septic re-inspections and for enhanced setbacks from surface water for new approvals.

6. Recommendations for Mitigation

Phosphorus loads from diffuse human sources represent a high proportion of the total load to Wasi Lake (32%) and Callander Bay (41%), which can be controlled by management practices (Figure 18).

Figure 18. Schematic illustrating phosphorus loads that can be controlled by management techniques.



The first step in developing a nutrient reduction strategy for a watershed is to identify the main nutrient sources in the watershed. This information is instrumental to developing watershed management measures that target the largest contributors and therefore will have the highest likelihood of improving downstream water quality. Based on this information, a priority list of issues to address can be developed that will guide the subsequent steps in the process, such as stakeholder consultation, financial considerations and decision making regarding the implementation of best management practices.

While the phosphorus budget for Callander Bay and Wasi Lake identifies agricultural practices and septic systems as the two primary human sources of phosphorus, additional information is required for both of these sources (see Section 5) to make definite management recommendations or to predict the potential loading reductions that could occur with implementation of management practices. At this stage, we provide a list of common Best Management Practices that focuses on agriculture and septic system sources (Table 14) as well as a list of “No Regrets” BMPs that will improve water quality, without the need for further investigation.

Table 14. List of Best Management Practices to Reduce Diffuse Phosphorus Sources in the Callander Bay Watershed

Category	Best Management Practices	Regulated (R)/ Voluntary (V)
Urban		
Stormwater Management	Source controls (pet waste collection, street cleaning, reduced fertilizer)	V
	Lot level controls (e.g. grading, infiltration, green roofs)	V
	Conveyance (transport) controls (permeable pavement, pervious pipe, grass swales)	V
	Stormwater treatment (e.g. constructed wetlands, sand filters, OGS ¹)	V
	Stormwater Ponds	V
	manufactured BMP systems (alum additions, etc)	V
Riparian	Buffer strips, riparian maintenance in urban areas	V
Agriculture		
Runoff from Crops	Match fertilizer application to crop nutrient requirements and soil properties	V
	Crop rotations	V
	Proper fertilizer application timing	V
	Cover crops during non-grow season	V
	Improved fertilizer storage	V
	Reduced or no tillage	V
	Buffer strips (Vegetated areas along waterways), riparian maintenance	R
	Irrigation management (e.g. low water-loss technologies, reduced system leakage, optimal irrigation timing)	V
Livestock Operation	Restrict livestock access to surface water	R
	Rotation of grazing pastures	
	Minimizing runoff from livestock yards	V
	Milkhouse wash water treatment	V
Runoff from Farm Yards	Stormwater retention ponds, constructed wetlands, berms (soil barrier), planted waterways etc.	V
Manure	Manure storage controls	V
	Manure treatment (dewatering & nutrient removal systems)	V
	Manure land application practices (e.g. crop requirements)	V
	Distance from waterways, buffer strips between piles and waterways	R
Airborne Nutrients	Wind breaks (trees, hedges etc. to reduce soil erosion)	V
Biosolids	Restrictions on timing of applications	R
	Setbacks, application factors (soil type, slope, compaction)	R
Shoreline Development		
Septic Systems	Design/installation and initial inspection	R
	Use of best available technology	V
	Maintenance - pump regularly etc.	V
	Follow-up inspections	V
	Use of phosphate free products (into septic)	V
	By-laws regulating new lot sizes in Official Plans	R
Overland Flow	Limit use of lawns and fertilizers	R/V
	Buffer strips, riparian maintenance	V
Recreation	Grey water (non-sewage wastewater) disposal from boats	V



All of the BMPs listed in Table 14 have the potential to reduce nutrient loads to surface and ground water in the Callander Bay watershed, but to a different degree based on the importance of the practice in the watershed and the percentage of nutrient removal by the specific BMP. In order to quantify the potential nutrient load reduction by a specific BMP, the size of the operations to be managed and the effectiveness of the BMP in terms of nutrient load reduction need to be determined. The effectiveness of BMPs in terms of percent reduction in nutrient loads has been estimated in a number of studies, particularly for agricultural BMPs, with results differing from study to study. Reviews of these studies have been completed, however, such that an approximate effectiveness for some BMPs can be provided (e.g., Table 15).

Table 15. Examples of Effectiveness of Best Management Practices for Phosphorus Load Reduction from Agriculture and Septic Systems

BMP Category	Specific BMP	Phosphorus Load without BMP	Phosphorus Export Reduction (%)
Milkhouse wastewater treatment	Flocculator	0.69 kg TP/cow/yr (excluding manure); 2.76 kg TP/cow/yr (including manure)	95-99
	Vegetated Filter Strip		7.2 – 100
	Settling Basins		5 – 67
	Constructed Wetland		45% - 99
	Anaerobic Lagoon		54-91
	Facultative Pond		5.5 – 91 (most > 80)
	Aerobic Lagoon		30-47
Manure storage	Daily Spreading	15.2 kg TP/cow/yr	90
	Dry & Roof		90
	Earthen		60-80
	Lagoon/flush		40-80
	Open Lot		70 +/- 20
	Pits & slats		95
	Scrape/storage tank		85-90
	Dairy pile manure		80
Clean water diversion	Roof Diversion for Feedlot manure	Same as for Manure Storage	70
	Roof Diversion for Stockpiled Dairy Manure		80
	Berm Diversion for Feedlot Manure		70 for portion of runoff that is being retained by berm (often ~half)
Restrict livestock access to streams	Fencing Off (providing alternative water source)*	0.46 kg/cow/yr (Beef) 0.23 kg/cow/yr (Dairy) (from manure only)	100 (effect on manure only)
	Fencing Off	Erosion loss to be calculated for access area	75 – 98 reduced TP loss from erosion
Conservation tillage	Disk	1 kg/ha/yr	93
	Ridge Till		59
	Reduced Till		85%
	No Till		61%
Buffer strips for streams through crop land	Width ≤ 5 m	1 kg/ha/yr	56%
	Width 6-10 m		67%
	Width 11 + m		74%
Cover crops			60%
Fragile land retirement			30%
Septic systems	Improve failing septic systems (only if within 50 m of a surface water body)	0.6 kg TP /capita/yr	70%

Notes: Source: South Nation Conservation 2003. "Phosphorus Loading Algorithms for the South Nation River". Updated Source Accounting Methodology for the Rural Water Quality Program (prepared by Chris Allaway, University of Ottawa).
*Providing alternative water source does not guarantee 100% reduction, but can still be effective (77% of reduction in stream bank loss and 98% in TP loading)



The phosphorus reduction rates presented in Table 15 demonstrate that BMPs provide large opportunities for reducing the phosphorus contribution of diffuse sources from agricultural lands and septic systems to Callander Bay.

Once data gaps have been filled for agricultural areas and septic systems, appropriate BMPs can be selected to best address phosphorus loadings from these sources.

Recommendation 5. “No Regrets” Mitigation

Although some specific recommendations for more information are presented above, there are Best Management Practices that produce environmental benefits that should be implemented to protect Callander Bay, even in the absence of complete documentation of their relative importance.

5.1 – Efforts to reduce the attractiveness of nearshore areas as habitat for Canada geese would reduce loading of phosphorus and bacteria to Callander Bay and its tributaries. This can be done by enhanced plantings of emergent vegetation in the littoral and riparian areas to discourage use by geese.

5.2 – Efforts to improve management of livestock and manure runoff to keep both away from surface waters provides immediate benefits in restoring riparian habitat (through elimination of grazing pressure and trampling) and reduced bacterial and nutrient loading. A survey of watershed streams should be undertaken, to identify candidate streams or stream areas for protection, as well as investigation of cooperative programs for fencing and riparian zone protection, and means to divert runoff away from manure piles to reduce the effectiveness of runoff as a pathway for loading to surface waters.

5.3 – Water quality and aquatic habitat in streams throughout the watershed would be improved by maintenance of riparian buffer strips of natural vegetation. These would shade streams, filter out particulate pollutants, take up dissolved nutrients and provide coarse particulate matter (fallen vegetation) for habitat, structure and carbon source to the streams. A survey of watershed streams is recommended to identify candidates for riparian enhancement programs.

5.4 Fertilizer applications to shoreside lawns are an unnecessary source of phosphorus load to surface waters. Stewardship initiatives to a) promote phosphorus free fertilizers or b) promote fertilizer-free lawns should be undertaken. A single application of 10:10:10 fertilizer to a 30m*30m lawn contains nearly 2 kg of phosphorus, and some of this may be mobilized to the water.

5.5 – Riparian buffer strips of natural vegetation provide habitat, filter particulate matter and take up dissolved nutrients. Naturalization of shorelines adjacent to Wasi Lake and Callander Bay should be promoted and enforced.

5.6 – Urban runoff contributes up to 500 kg/yr, (8%) of the phosphorus load to Callander Bay. This represents ~20% of the human source. An investigation of stormwater pathways to Callander Bay should be undertaken, and a catchment by catchment survey of potential means to reduce them by promoting infiltration, stormwater detention, sheet flow through grassy swales and reductions in urban fertilizer use.



5.7 – Any future changes in land use (development) should adopt BMPs to reduce the likelihood and magnitude of phosphorus runoff. All development applications within the watershed should be reviewed to ensure that the potential for phosphorus runoff has been minimized.

7. Implications for Source Protection Planning

The Technical Rules for Drinking Water Source Protection under the Clean Water Act (2002) require the delineation of an ‘Issue Contributing Area (ICA)’, that is, the area within which activities contribute to the concentration of a contaminant at a drinking water intake that is listed as a drinking water issue. For the Callander intake technical studies completed by HESL (2010), phosphorus was listed as a drinking water issue based on the documented occurrence of toxin-producing cyanobacteria blooms in Callander Bay and the known relationship between phosphorus concentrations and algal bloom activity. Recognizing that there are natural sources of phosphorus in the watershed and without knowledge of the contribution of phosphorus from human activities, the ICA was defined as the entire vulnerable area of the Callander intake (i.e., all IPZ areas), which is the maximum area allowed by the Technical Rules. Uncertainty remained, however, in the representativeness of the defined ICA to capture the primary sources of phosphorus to Callander Bay from human activities.

Results of the phosphorus budget for Callander Bay indicate that a large portion of the land area in the Callander Bay watershed is encompassed by the Intake Protection Zones, or the Issue Contributing Area (ICA). Furthermore, human sources of phosphorus in the ICA contribute a large portion (up to 84%) of the loading from human sources to Callander Bay. Based on these findings, it is concluded that the ICA defined by HESL (2010) does capture the primary sources of phosphorus to Callander Bay from human activities in the watershed and recommend that the ICA remain as defined.

In addition to the delineation of the ICA for the Callander intake, the phosphorus budget can be used to better inform the classification of threats (i.e., as significant, moderate or low) for drinking water source protection. Threats are defined by the Technical Rules as activities that contribute or potentially contribute to a contaminant at the intake. For threats related to phosphorus, this could be achieved by ranking human phosphorus sources according to their potential phosphorus loading contribution. This would, however, require refinements of the phosphorus budget to specifically account for loadings from different agricultural activities and from septic systems.

8. Conclusions

The phosphorus budget for Callander Bay derived from export coefficient modelling and measured phosphorus loads provides a reasonable estimate of phosphorus loadings from all major sources in the watershed, including natural sources (i.e., atmospheric deposition and runoff from undisturbed land areas) and human sources (i.e., agriculture, urban runoff, septic systems and STP effluent). Validation of the phosphorus budget with measured phosphorus



loads in the Wasi River and by comparison of measured and modelled phosphorus concentrations in Wasi Lake provides a high degree of confidence in the total load estimates. Uncertainty in the relative contribution of phosphorus loadings from septic systems and different agricultural practices remains, however, but can be addressed with the collection of additional site-specific information.

Human sources account for approximately 43% of the total phosphorus loading to Callander Bay and 32% of the loading to Wasi Lake, a large portion of which can be controlled by Best Management Practices. Identification of the most appropriate BMPs for Callander Bay and Wasi Lake requires refinement of the phosphorus budget to better account for loadings from different types of agricultural practises and from septic systems. Nevertheless, a series of “No Regrets” BMPs can be implemented at low cost, and these will improve water quality and produce other benefits.

While considerable load reductions can be achieved by BMPs, the natural phosphorus loading to Callander Bay and Wasi Lake is large such that these water bodies may remain relatively productive with potential for algal bloom activity even if all human sources of phosphorus were eliminated. In particular, phosphorus concentrations in Callander Bay increased in the years after 1950, when the Portage Dam was built at the outlet of Lake Nipissing. Nevertheless, phosphorus load reductions and resultant reductions in phosphorus concentrations in Callander Bay and Wasi Lake over current levels would reduce the risk of cyanobacteria blooms.

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10. Appendices

Phosphorus and hydrologic data files (digital)

